Atmos. Meas. Tech. Discuss., 7, 5953–6019, 2014
www.atmos-meas-tech-discuss.net/7/5953/2014/
doi:10.5194/amtd-7-5953-2014

Atmospheric Measurement Techniques Discussions

A gas chromatograph for quantification of peroxycarboxylic nitric anhydrides calibrated by thermal dissociation cavity ring-down spectroscopy

T. W. Tokarek, J. A. Huo, C. A. Odame-Ankrah, D. Hammoud, Y. M. Taha, and H. D. Osthoff

Department of Chemistry, University of Calgary, Calgary, Alberta T2N 1N4, Canada

Received: 16 April 2014 – Accepted: 29 May 2014 – Published: 13 June 2014

Correspondence to: H. D. Osthoff (hosthoff@ucalgary.ca)

Published by Copernicus Publications on behalf of the European Geosciences Union.

Abstract

The peroxycarboxylic nitric anhydrides (PANs, molecular formula $RC(O)O_2NO_2$) can readily be observed by gas chromatography coupled to electron capture detection (PAN-GC). Calibration of a PAN-GC remains a challenge because the response factors (RF's) differ for each of the PANs and because their synthesis in sufficiently high purity is non-trivial, in particular for PANs containing unsaturated side chains. In this manuscript, a PAN-GC and its calibration using diffusion standards, whose output was quantified by blue diode laser thermal dissociation cavity ring-down spectroscopy (TD-CRDS), are described. The PAN-GC peak areas correlated linearly with total peroxy nitrate (PN) mixing ratios measured by TD-CRDS ($r > 0.96$). Accurate determination of RF's required the concentrations of PAN impurities in the synthetic standards to be subtracted from PN. The PAN-GC and its TD-CRDS calibration method were deployed during ambient air measurement campaigns in Abbotsford, BC, from 20 July to 5 August, 2012, and during the Fort McMurray Oil Sands Strategic Investigation of Local Sources (FOSSILS) campaign at the AMS13 ground site in Fort McKay, AB, from 10 August to 5 September 2013. For the Abbotsford data set, the PAN-GC mixing ratios were compared and agreed with those determined in parallel by thermal dissociation chemical ionization mass spectrometry (TD-CIMS). Advantages and disadvantages of the PAN measurement techniques used in this work and the utility of TD-CRDS as a PAN-GC calibration method are discussed.

1 Introduction

Peroxycarboxylic nitric anhydrides (PANs, also referred to as peroxyacyl nitrates) are important trace gas constituents of the troposphere (Roberts, 1990, 2007; Altshuller, 1993). PANs are formed as byproducts in the same photochemistry between NO_x and volatile organic compounds (VOCs) that produces ozone (O_3) and photochemical smog; they are thus good chemical markers of the types of VOCs involved in the

AMTD
7, 5953–6019, 2014

O_3-formation process (Roberts, 2007). Figure 1 shows examples of common PANs that have been found in the troposphere. Peroxyacetic nitric anhydride (PAN, molecular formula $CH_3C(O)O_2NO_2$, which is often referred to by its non-IUPAC name peroxyacetyl nitrate) is generally the most abundant PAN species, because its main precursor, acetaldehyde, is an oxidation product of many VOCs. Peroxypropionic nitric anhydride (PPN) is derived from propanal and generally regarded as an anthropogenic pollution tracer, whereas peroxymethacrylic nitric anhydride (MPAN), which in ambient air is formed from methacrolein (an isoprene oxidation product), is a tracer of the oxidation of biogenic VOCs. Peroxyacrylic nitric anhydride (APAN, originally referred to as vinyl-PAN, Grosjean et al., 1994) is formed from acrolein (a propylene or butadiene oxidation product), and regarded as a tracer of petrochemical VOC oxidation chemistry (Roberts et al., 2001, 2003; Tanimoto and Akimoto, 2001). In high concentrations, PANs are lachrymators and toxic to plants (Horvath et al., 1986; Kleindienst, 1994). Furthermore, PANs are considered NO_x reservoir species because they are longer-lived than NO_x ($\equiv NO + NO_2$) and generate NO_2 when they decompose (Kirchner et al., 1999). Chemical transport models indicate that PANs can carry NO_x over long distances and affect photochemical O_3 production in remote regions (Fischer et al., 2010; Pandey Deolal et al., 2013). Due to its relatively long lifetime, PAN is also often the most abundant "odd nitrogen", or NO_y ($\equiv NO_x + \text{PAN} + \text{AN} + HNO_3 + HONO + NO_3 + 2N_2O_5 + ClNO_2 + \ldots$), species in the mid to upper troposphere, at high latitudes, and in biomass burning plumes (Stroud et al., 2003; Alvarado et al., 2010).

Over the years, a variety of techniques to quantify PAN mixing ratios in ambient air have been developed. A widely used, mature, robust, and sensitive technique is gas chromatography with capillary columns and electron capture detection (e.g., Darley et al., 1963; Blanchard et al., 1990; Schrimpf et al., 1995; Williams et al., 2000; Volz-Thomas et al., 2002; Flocke et al., 2005); this measurement technique is referred to as PAN-GC in this manuscript. A relatively new and increasingly popular technique is thermal dissociation chemical ionization mass spectrometry (TD-CIMS). In this technique, the PANs are dissociated in a heated inlet to their corresponding peroxyacyl

AMTD
7, 5953–6019, 2014

radicals and detected, following electron transfer reaction with iodide reagent ion, as their corresponding carboxylate anions (Slusher et al., 2004). Measurement of PANs using TD-CIMS in ambient air, however, is not straightforward as there are matrix effects and interferences to consider (Zheng et al., 2011; Mielke and Osthoff, 2012; Phillips et al., 2013). Accurate measurements of MPAN are particularly challenging because TD-CIMS is relatively insensitive to MPAN and because ions other than methacrylate (CH_2=C(CH_3)CO_2^-), e.g., trifluoromethoxy anion (CF_3O^-, generated from CF_3OH that outgasses from Teflon tubing) and crotonate (CH_3CH=$CHCO_2^-$, generated from peroxycrotonyl nitrate, CPAN), interfere at *m z* 85 (Zheng et al., 2011; Mielke and Osthoff, 2012). PANs have also been quantified by proton-transfer mass spectrometry (Hansel and Wisthaler, 2000; Hastie et al., 2010) and by quantifying the NO_2 generated from their thermal dissociation (usually by difference relative to background NO_2) using luminol chemiluminescence (CL) (Nikitas et al., 1997; Marley et al., 2004), laser-induced fluorescence (LIF) (Day et al., 2002; Wooldridge et al., 2010), or cavity ring-down spectroscopy (CRDS) (Paul and Osthoff, 2010; Paul et al., 2009). In spite of the development of these new PAN measurement techniques, PAN-GCs remain attractive because of their relative simplicity, compactness, robustness, good sensitivity, and lack of interferences, properties which are desirable for long-term and unattended operation (e.g., Mills et al., 2007; Zhang et al., 2009, 2012; Fischer et al., 2010).

The key challenge in the measurement of PANs by PAN-GC is calibration. The response factors (RF's) of PAN-GC instruments (and, incidentally, of TD-CIMS instruments) are different for each of the PAN compounds; hence, each PAN requires individual calibration (Flocke et al., 2005). Diffusion sources whose outputs are quantified by NO_y CL instruments are commonly used for this purpose. This, however, requires syntheses of PAN standards in high purity. The syntheses of pure MPAN and APAN samples are particularly challenging, because both molecules contain carbon-carbon double bonds in their side chain which are prone to unwanted side reactions. In their case, a preparative-scale GC is often used to discriminate the PAN standard from other nitrogen oxide containing impurities (Flocke et al., 2005). Another challenge with the

AMTD
7, 5953–6019, 2014

calibration using diffusion sources is sample storage and delivery, as PANs are explosive in pure form and are prone to thermal decomposition; as a result, standards are usually stored in a non-polar solvent (such as tridecane) at cold temperatures, e.g., in vials partially immersed in an ice-water bath (Gaffney et al., 1984). Photochemical sources, in which PANs are generated in situ, e.g., from photodissociation of dialkyl ketones in the presence of oxygen and a calibrated amount of NO_x, are an attractive alternative to diffusion standards as they require neither cooling nor the transport and storage of toxic chemicals (Warneck and Zerbach, 1992; Grosjean et al., 1984; Flocke et al., 2005; Pätz et al., 2002; Volz-Thomas et al., 2002; Emrich and Warneck, 2000; Furgeson et al., 2011). So far, photochemical sources have been limited to the generation of PANs with short saturated side chains, i.e., PAN, PPN, and peroxyisobutyric nitric anhydride (PiBN) (Furgeson et al., 2011).

In this manuscript, a recently assembled PAN-GC and its calibration for PAN, PPN, MPAN, and APAN using diffusion standards whose outputs were quantified by blue diode laser TD-CRDS are described. Calibration and ambient air data collected during measurement campaigns at ground sites in Abbotsford, BC, from 20 July to 5 August 2012, and near Fort McKay, AB, during the Fort McMurray Oil Sands Strategic Investigation of Local Sources (FOSSILS) campaign from 10 August to 5 September 2013, are presented.

2 Experimental

2.1 PAN-GC setup and operation

The PAN-GC used in this work was a Varian 3380CP equipped with an ECD detector (Model number: 02 001972 01) which was converted to measure PANs in a similar fashion as described by Fischer et al. (2010). A 10-port, 2-position sample valve (Vici Valco EHC10WE, 1 16 ×0.40 mm with microelectric actuator) and a sample loop constructed from polyether ether ketone (PEEK, VICI "Cheminert") were mounted inside

AMTD
7, 5953–6019, 2014

the GC oven as shown in Fig. 2. The analytical column used was a medium polarity "megabore" column (Restek RTX-1701) with 0.53 mm inner diameter (i.d.) and 1 µm film thickness. Sample air was drawn through the sample loop using a miniature air compressor (McMaster-Carr 4404K15) at a flow rate of 115 standard cubic centimeters per minute (sccm) set by flow restriction. Ultrapure He (Scott-Marrin) was used as carrier gas at a flow rate of 20 mL min^{-1} set using a needle valve and measured at the ECD exhaust. A van Deemter plot analysis (not shown) showed that the chromatographic height of a theoretical plate was at its minimum value at this carrier gas flow rate. The He regulator was connected to the GC via PEEK 1 8 outer diameter (o.d.) tubing (VICI) through a He-specific inline "triple trap" (Restek 22469) to remove hydrocarbons, oxygen, and moisture. The carrier gas was re-humidified by passing it over 50 g $CuSO_4$ pentahydrate (99.995 %, Aldrich) contained in a stainless steel sample cylinder with 150 cm^3 internal volume capped by a 2 µm pore size in-line filter (Swagelok 316L-HDF4-150 and SS-4FW5-2). The presence of H_2O has been reported to reduce chemical on-column losses of PAN compounds and to improve run-to-run reproducibility (Flocke et al., 2005). ECD-grade N_2 (Praxair) was used as make-up gas at a flow rate of 2.0 mL min^{-1}.

Prior to making measurements, the ECD was heated for 24 h at its maximum temperature setting of 300 °C under He flow to lower the background signal. Afterwards, the ECD was operated at its lowest possible temperature setting (50 °C) in constant current mode with the "FAST" (50 ms) time constant setting. The ECD signal voltage was digitized using a Universal Serial Bus (USB) 2.0 data acquisition module (Omega) connected to a laptop computer running software written in National Instrument's Labview 2010, which also controlled the main switching valve. The valve was set to the "fill sample loop" position (Fig. 2a) by default. Injections were automated (usually every 6 or 10 min) synchronized to the top of the hour by switching to the "sample to column" setting (Fig. 2b) for 30 s.

During the Abbotsford campaign, the PAN-GC sampled from an inlet constructed from 1 4 (0.635 cm) o.d. fluorinated ethylene propylene (FEP) Teflon tubing shared

AMTD
7, 5953–6019, 2014

with a commercial O_3 UV absorption spectrometer (Thermo 49i) and sampled through a 1 µm pore size Teflon filter in a polycarbonate filter holder (Pall Life Sciences). The passing efficiency of such a filter has been reported as 100 % (Roberts et al., 2007), which was confirmed in laboratory tests. Ambient air samples were automatically injected onto the column once every 6 min. The sample had an internal volume of 2.0 mL, and the analytical column length was 20.41 m. The column oven was set to a temperature of +25 °C, but was occasionally higher in the afternoons as the trailer housing the instruments warmed up above this temperature. At the beginning of the campaign (until 25 July, 16:00 local time), the EC was operated with the "CAP" (480 pA) sensitivity setting and at a contact potential of 175 mV. The remaining data were collected using the (less sensitive) "N_2STD" (290 pA) setting. The detector voltages were digitized using a 14-bit data acquisition module (Omega USB-1408FS) at a sampling rate of 500 Hz.

For the FOSSILS campaign, the data acquisition board was upgraded to a 16-bit module (Omega USB-1608FS) operated at a sampling rate of 10 kHz. Data were boxcar averaged to 20 Hz immediately after acquisition. The ECD was operated with the "CAP" (480 pA) sensitivity setting at a contact potential of 400 mV. The analytical column was shortened to 10.5 m, the sample loop volume reduced to 1.0 mL, and the oven temperature set point was set to a temperature of 22 °C. Ambient air samples were automatically injected once every 10 min. Partway through the FOSSILS campaign, the PAN-GC's miniature air compressor pump failed. Sample air was then drawn through the sample loop using a 50 µm critical orifice connected to a twin head diaphragm pump (KNF Neuberger N 026.1.2AT), which provided a flow rate of 15 sccm as measured using a mass flow controller (MKS M100B) and altered the PAN-GC RF's (see Sect. 3.3).

AMTD
7, 5953–6019, 2014

2.2 Data reduction

To determine chromatographic parameters, parameters of Gaussian Eq. (1) were fitted to the observed peaks using a macro written in Igor Pro 6 (Wavemetrics).

$$V = \frac{A}{\sqrt{2\pi\sigma^2}} e^{\frac{-1}{2}\left(\frac{t-t_0}{\sigma}\right)^2} + V_0 + mt \quad (1)$$

In Eq. (1), V is the signal voltage observed, A and σ are the peak area and its standard width, t is time since injection, t_0 the retention time, and V_0 and m are the offset and slope of the background.

Chromatograms were smoothed in post-data acquisition processing using Igor Pro's built-in Savitzky-Golay algorithm (4th order). For the Abbotsford campaign, a width of ~ 2 s, i.e., 1001 data points, was chosen, whereas for the FOSSILS campaign a width of four standard peak half widths (~ 4σ, see Tables 1 or 2 for typical σ values) was used for each peak, i.e., PAN peaks were smoothed over 69 data points, PPN over 235 points, and MPAN over 399 points.

To ensure that the fitting macro converged onto the "correct" local minimum (i.e., fitted the peak and did not attempt to fit a trend in the background or the oxygen peak), only the region near a peak ($t_0 \pm 8\sigma$) was fitted. The fits were performed in stages: the first iteration constrained t_0, σ, and A to values from the preceding chromatogram (or to values that were manually entered) to determine the baseline (which shifted often between chromatograms). These constraints were sequentially lifted in subsequent fits until the peak was accurately described by Eq. (1). For PPN and MPAN, knowledge of the relative retention times and peak widths from calibration runs were used in the first fit iteration, i.e., t_0 and σ were estimated from their average values relative to those of the PAN peak (in the same chromatogram) observed in calibration runs. Subsequent fit iterations were the same as for PAN.

AMTD
7, 5953–6019, 2014

2.3 Synthesis of standards

PAN standards were synthesized following procedures communicated to us by J. Roberts (personal communication, 2009) that are based on those described in earlier publications (Williams et al., 2000; Bertman and Roberts, 1991). Reagents were purchased from Sigma-Aldrich and used as received. PAN, PPN, and MPAN were synthesized from their respective acid anhydrides, which were oxidized using H_2O_2 to the peroxyacids and reacted with strong sulfuric and nitric acid. Workup and sample storage were as described earlier (Mielke and Osthoff, 2012). APAN was synthesized from acryloyl chloride ($CH_2CHC(O)Cl$) as it was commercially available whereas acrylic anhydride was not. Briefly, 8.0 mL of acryloyl chloride ($CH_2CHC(O)Cl$) were placed into an ice-cold 100 mL round-bottom flask containing a magnetic stirrer, to which 5.0 mL of cold 50 % H_2O_2 were added drop wise. After 2 h, 20.0 mL of cold tridecane were added, followed by slow addition of 5.0 mL of cold concentrated sulfuric acid and slow drop wise addition of 6.0 mL cold concentrated nitric acid. After 15 min, the organic and aqueous layers were separated in a 125 mL separatory funnel. The organic layer was washed three times using 50 mL of cold deionized water, dried over $MgSO_4$, and filtered through cotton swab. Small aliquots of APAN in tridecane solution were stored in 2.0 mL polypropylene centrifuge tubes (VWR) in a freezer prior to use.

2.4 Thermal dissociation cavity ring-down spectroscopy and chemical ionization mass spectrometry

The University of Calgary blue diode laser TD-CRDS has been described earlier (Paul and Osthoff, 2010). Briefly, the TD-CRDS contains multiple sample channels operated at different inlet temperatures and differentiates between NO_2 sampled at ambient temperature and NO_2 + total peroxyacyl nitrates (PN) sampled at an inlet temperature of 250 °C. NO_2 is quantified by its optical absorption at 405 nm. PN is quantified by difference relative to the ambient temperature channel. During the FOSSILS campaign, the blue diode laser TD-CRDS was operated with the "linear regression of the sum"

AMTD
7, 5953–6019, 2014

(LRS) algorithm (Everest and Atkinson, 2008; Taha et al., 2013) and a diode laser pulse repetition rate of 1 kHz.

The iodide ion TD-CIMS and its operation have also been described in earlier publications (Mielke et al., 2011, 2013; Mielke and Osthoff, 2012). During the Abbotsford campaign and in laboratory experiments, it was operated in TD mode (i.e., with an inlet heated to 190 °C) with a humidified reaction chamber. Background counts arising from peroxyacids (Furgeson et al., 2011; Phillips et al., 2013) were determined by periodically (every 30 min) flooding the inlet with a high concentration of NO delivered from a 0.1 % NO in N_2 cylinder (Praxair). The background count data were interpolated and subtracted from the ambient ion counts. The TD-CIMS method is prone to interferences arising from proton exchange of acetate (which is generated from PAN) with organic acids and from titration of the peroxyacyl radicals by NO and NO_2 (Mielke and Osthoff, 2012). In previous work, the titration by NO_x was tracked by adding ^{13}C-PAN as internal standard to the inlet. However, the photochemical source that had been used in earlier experiments to deliver ^{13}C-PAN was not used in Abbotsford because it had been contaminated (with Cl_2) immediately prior to the campaign. To account for matrix effects arising from titration of the peroxyacyl radicals by NO and NO_2 in the field, ambient ion counts were multiplied by a factor of $\exp(k_{NO} \times [NO] + k_{NO_2} \times [NO_2])$, where [NO] and $[NO_2]$ are mixing ratios observed by a commercial NO/NO_y CL instrument (Thermo 42i) and the University of Calgary CRDS, respectively, and k_{NO} and k_{NO_2} are empirical constants. The values for PAN ($k_{NO} = 0.0513\,\text{ppbv}^{-1}$ and $k_{NO_2} = 0.0203\,\text{ppbv}^{-1}$) were determined from a comparison of our group's TD-CIMS PAN data with PAN measured by NOAA's PAN-GC during the Calnex-LA campaign (Mielke et al., 2013). The values for PPN ($k_{NO} = 0.118\,\text{ppbv}^{-1}$ and $k_{NO_2} = 0.033\,\text{ppbv}^{-1}$) were taken from Mielke and Osthoff (2012). During FOSSILS, the CIMS was operated with an ambient-temperature inlet, and PANs were not quantified by TD-CIMS.

AMTD
7, 5953–6019, 2014

2.5 Inlet setup and in-field calibrations

Figure 3 shows a sketch of the inlet setup deployed in the field. All instruments sampled from a common inlet manifold and were "zeroed" periodically using automated switching valves, usually in between PAN-GC injections. During the Abbotsford campaign, a flow of ~20 standard liters per minute (slpm) of ultrapure ("zero" grade) air was delivered from gas cylinders (Praxair) via a 30 slpm capacity mass flow controller (MFC, Celerity) and a 2-way normally closed solenoid valve (Galtek 203-1414-215). During FOSSILS, the instruments were zeroed using a combination of ambient and exhaust air that was scrubbed with a combination of activated charcoal (Whatcom) and potassium permanganate (Purafil Select). Occasionally, zero air from a cylinder was used instead. The only difference observed between the scrubbed air and the zero air was the higher humidity of the former.

The PAN-GC was calibrated using diffusion sources that each contained a PAN standard of interest whose output was quantified by TD-CRDS. For in-field calibration runs, the gas stream containing the PAN standard was delivered to the inlet manifold in a similar fashion as described for N_2O_5 by Odame-Ankrah and Osthoff (2011). Briefly, a 1 4 o.d. FEP Teflon tube was conditioned for 30 min to 1 h with the output of the diffusion standard (which was kept in a glass vessel in an ice-water bath) diluted in typically 20 sccm zero air. This tube was tee'd into the line delivering zero (or scrubbed) air to the main inlet (Fig. 3) and connected, via a remotely controlled normally open 2-way valve (Cole Parmer 01540-10), to a waste line evacuated by a diaphragm pump (KNF Neuberger N026.1.2AT) whose flow rate was throttled using a 50 µm critical orifice (Lennox Laser). A similar setup was also used to deliver a flow containing high concentrations of NO to zero the TD-CIMS.

A typical sequence of how an individual calibration point was obtained is shown in Fig. 4. Prior to and following the PAN-GC injection, the TD-CRDS and NO/NO_y CL instruments were zeroed (shown with a gray underlay) by flooding the inlet with scrubbed or zero air. About 2 min prior to the GC injection, the standard was delivered to the

AMTD
7, 5953–6019, 2014

instruments by closing the 2-way valve. Following the PAN-GC injection, the other instruments were zeroed again. The bracketing of the calibration sequence by "zeroes" is required to make accurate measurements by TD-CRDS, which is inherently a measurement of difference, i.e., of ring-down time constants in the absence and presence of the optical absorber (NO_2 in this case).

During the Abbotsford campaign, a PAN calibration point was obtained typically once (sometimes twice) per day (for a total of 16 points). In addition, 6 calibration points were collected for PPN. During FOSSILS, calibration points were collected approximately daily for PAN (34 points in total), PPN (21 points), and MPAN (26 points). The PAN standards were kept in separate glass traps which were stored in ice/water baths between uses. Each diffusion source was loaded only once at the beginning of each campaign, and there was no evidence for sample decomposition. The RF for APAN was determined after the FOSSILS campaign.

The NO/NO_y CL instrument was calibrated daily (against CRDS) by sampling NO and/or NO_2, generated using an NO standard gas cylinder (Scott Marrin) to which up to a stoichiometric amount of O_3 was added, generated by passing oxygen from a gas cylinder (Praxair) past a 254 nm Hg lamp. The NO_x standard was delivered to the main inlet via a separate 1 4 o.d. Teflon tube connected to the zero air line; delivery was automated using a 2-way normally open valve.

2.6 Ambient air measurement locations

The first set of ambient air measurements was conducted at a routine monitoring site located east of the Abbotsford International Airport (International Air Transport Association (IATA) airport code: YXX) at latitude 49.0212° N and longitude −122.3267° W, ∼ 60 m above sea level (a.s.l.). The City of Abbotsford (population ∼ 140 000) is located in the Lower Fraser Valley which is prone to episodes of poor air quality, i.e., exceedances of O_3 and particulate matter (PM) air quality standards. These occur in part because the topography facilitates stagnation periods and trapping of airborne pollutants near the surface and because of continuously increasing emissions from urban,

AMTD
7, 5953–6019, 2014

suburban, agricultural and marine sources that include the City and Port of Vancouver (Vingarzan and Li, 2006; Ainslie and Steyn, 2007). The measurement site itself was surrounded by raspberry fields and impacted by emissions from nearby agricultural operations (such as chicken farms), and traffic on nearby highways and secondary roads.

The second set of ambient air measurements (the FOSSILS campaign) was conducted at the AMS13 (formerly "Syncrude") ground site near Fort McKay, AB, at latitude 57.1492° N and longitude −111.6422° W and ∼ 270 m a.s.l. This site is surrounded by boreal forest and is occasionally impacted by emissions from nearby oil sands surface mining operations, for example, the Syncrude mine ∼ 4.5 km to the South, the Suncor mine ∼ 20 km to the SE, Canadian Natural Resources Ltd. (CNRL) operations ∼ 13 km to the NNW, and the Shell mine ∼ 8 km to NNE.

3 Results

3.1 Gas chromatography of PAN standards

Figure 5a shows sample chromatograms of PAN, PPN, and APAN standards plus a mixture of all three (all in zero air) collected in the laboratory after the FOSSILS campaign. Figure 5b shows sample chromatograms of PAN, PPN, and MPAN standards (in a background of scrubbed air) acquired in the field during the FOSSILS campaign. Following the initial off-scale O_2 peak, the four PAN standards elute as base-line resolved peaks with relative retention times consistent with those reported by Tanimoto and coworkers for a RTX-1701 column (Tanimoto et al., 1999; Tanimoto and Akimoto, 2001). Tables 1 and 2 summarize the fits of parameters in Eq. (1) to the chromatograms presented in Fig. 5. A comparison of the fits and observations are shown as inserts in Fig. 5. The PANs eluted as essentially perfect Gaussians. As a result, the retention times, the peak widths, and the areas were determined with high precision. In contrast, peaks not associated with PANs (i.e., alkyl nitrates and other impurities) exhibited tailing and were not quantified in this work.

AMTD
7, 5953–6019, 2014

3.2 Impurities in the PAN standards

The chromatograms in Fig. 5 show that the PPN, MPAN, and APAN standards each contained a small amount of PAN. This impurity was observed in all synthetic batches analyzed in this work. The chromatogram of APAN in Fig. 5a and to lesser extent that of MPAN in Fig. 5b contained PPN as a result of memory effects within the calibration line; we found that PPN strongly absorbed and only very slowly desorbed from Teflon tubing. One Teflon fitting exposed to high PPN concentrations emitted PPN months after the exposure, in spite of several hour-long wash cycles in sonicated hot water. In the field, this memory effect was minimized (but could not be completely eliminated) by flushing the calibration tubing with 20 sccm zero air between calibrations. Because all PANs contribute to the PN signal observed by TD-CRDS, their mixing ratios need to be subtracted for accurate calibrations (see Sect. 3.3).

The standard chromatograms also contained traces of other impurities. For example, the shoulder on the O_2 peak in the PAN standard chromatogram (blue trace in Fig. 5b, retention time 41 s; not resolved in Fig. 5a) is likely methyl nitrate. A tailing peak in the PPN standard chromatogram (retention time of 51 s in the red trace in Fig. 5a; 71 s in Fig. 5b) is likely ethyl nitrate as it has the same retention time as an authentic ethyl nitrate sample.

The APAN standard chromatogram (purple trace in Fig. 5a) contains several unidentified peaks with retention times at 39 s, 73 s, 83 s, 122 s, 159 s, 242 s, and 284 s. To gain insights into the identity of these unknown peaks, the iodide ion chemical ionization mass spectra of an APAN standard (Fig. 6, red color) and of background room air (Fig. 6, blue color) were recorded. The acrylate anion originating from APAN appears at $m\ z$ 71 as the dominant product ion in the mass spectrum. In addition, nitryl chloride ($ClNO_2$) is observed as its iodide cluster ion at $m\ z$ 208 and $m\ z$ 210 and as its characteristic fragments at $m\ z$ 162 and $m\ z$ 164 (ICl^-) and $m\ z$ 35 and $m\ z$ 37 (Cl^-) (McNeill et al., 2006; Mielke et al., 2011; Osthoff et al., 2008). The amount of $ClNO_2$ evaporating from the tridecane solution was greater than that of APAN (1.5 ppbv vs.

AMTD
7, 5953–6019, 2014

0.8 ppbv), such that its presence would pose a significant interference if the instruments had been calibrated using NO_y CL. The APAN standard mass spectrum also contains $m\ z$ 59 (acetate, produced from PAN) as the third most abundant compound (~ 9 % relative to APAN), which is consistent with the GC chromatogram in Fig. 5a. In addition, there are enhancements at $m\ z$ 62 (nitrate) and smaller enhancements at $m\ z$ 73 (propionate, from PPN, ~ 7 % relative to APAN), $m\ z$ 75 (possibly $HOCH_2COO^-$ which is the anion derived from $HOCH_2COONO_2$, HPAN, ~ 3 % relative to $m\ z$ 71), $m\ z$ 105, $m\ z$ 107, and $m\ z$ 160 (unidentified). The relative peak intensity of the peaks at $m\ z$ 105 and $m\ z$ 107 suggests a Cl containing product, possibly $ClCH{=}CHCOO^-$, which may have been generated from $ClCH{=}CHCOONO_2$. The ion counts at $m\ z$ 105 are ~ 2 % relative to $m\ z$ 71.

3.3 PAN-GC response factors

The RF's of the PAN-GC were sufficiently stable between days such that conventional calibration plots, i.e., plots of PAN-GC peak areas against PN mixing ratios, could be constructed using all calibration points collected. The calibration plots were linear ($r > 0.96$ in all cases, and often > 0.99) and allowed precise determination of the RF's which are summarized in Tables 3 and 4. Alternatively, RF's could have been calculated from the daily ratios of GC area against PN; however, since this method would only be a single-point calibration, it was judged to be less accurate and reliable than the conventional calibration plot method.

Table 3 summarizes the RF's determined during the Abbotsford campaign. For PAN, the initially used "CAP" setting was more sensitive (by a factor of 2.8) than the N_2 STD setting but was less reproducible (5 % relative standard deviation (RSD) vs. 2 % RSD). The lower reproducibility was coincidental and likely caused by greater temperature fluctuations within the trailer during the initial days of the measurements (the PAN-GC was next to a door). Six PPN calibrations were also performed which showed PPN to have about the same RF as PAN (0.99 ± 0.04).

AMTD
7, 5953–6019, 2014

Figure 7 shows the calibration plots for the FOSSILS campaign. The output of the diffusion sources was deliberately varied between calibration runs during FOSSILS; as a result, the calibration data cover a wide range of mixing ratios (from ~ 0.2 ppbv to ~ 10 ppbv). Tables 4 and 5 summarize the RF's determined during the FOSSILS field campaign. Each RF was determined with a very low relative standard deviation (RSD, < 5 %, and often < 2 %).

At times substantial corrections (in particular for MPAN) had to be applied to the PN data because the calibration standard contained one or more of the other PANs as an impurity. During the FOSSILS campaign, the purity of the MPAN diffusion standard improved over time (from ~ 73 % PAN content at the beginning of the campaign to 6 % near the end), which suggests that PAN (the major impurity in the MPAN standard) is more volatile than MPAN.

The RF's when the PAN-GC was operated with a slower sample flow rate (Table 5, Fig. 7b) were systematically greater than those when the PAN-GC was operated with a faster sample flow rate (Table 4, Fig. 7a). In addition, the standard widths of the fits were ~ 20 % larger (Fig. A1). The differences arise from a change in the absolute number of molecules injected onto the column; while the sample loop volume (1.0 mL internal volume) was the same in both cases, the pressure inside the sample loop was lower at a faster flow rate because the 1 16 o.d. connecting tubing acted as a flow restriction.

Table 6 and Fig. A2 summarize PAN-GC RF's of PAN, PPN and APAN determined in the laboratory following the FOSSILS campaign. The relative RF's of PPN : PAN were $(0.96 \pm 0.04) : 1$, slightly higher than during FOSSILS and slightly lower than in Abbotsford. The APAN peak areas linearly correlated with the PN mixing ratios ($r = 0.981$). As shown in Sect. 3.2, the APAN standard contained several PAN-type compounds as impurities; however, only the mixing ratio of PAN could be accurately determined and subtracted from PN. The RF for APAN is hence a lower limit to its true value. The sum of impurity peak counts (other than those of PAN) observed in the CI mass spectrum (Sect. 3.2) is approximately 12 % relative to APAN. Conservatively assuming equal

AMTD
7, 5953–6019, 2014

sensitivity of the TD-CIMS to all PANs, the RF of the PAN-GC for APAN is estimated at 0.85 relative to PAN. If the TD-CIMS sensitivity to the impurities is lower than to PAN (which is likely, see Sect. 3.4), the PAN-GC RF relative to PAN approaches unity.

3.4 TD-CIMS response factors

The output of the diffusion sources was also used to determine the RF's of the TD-CIMS, for PAN and PPN in Abbotsford (Table 3) and for all four PANs in laboratory experiments (Fig. A3 and Table 8). The RF's for PAN, PPN, and MPAN are consistent with those reported earlier by our group with the same instrument (Mielke et al., 2011, 2013; Mielke and Osthoff, 2012) and by Zheng et al. (2011) (Table 9). For MPAN, some variability in the RF is expected as the MPA radical dissociates at (and at a rate that varies sharply with) the TD-CIMS inlet temperature. The RF for APAN, corrected for ~ 21 % total impurities (Sect. 3.2), was > (0.75 ± 0.02), and likely higher still if the sensitivity of the impurities is less than that of PAN. The range of values is greater in this work than reported by Zheng et al. (2011), who reported obtaining different RF's depending on calibration method, interpreted as caused by standard impurities such as acrylic acid.

3.5 Measurement of PAN and PPN in Abbotsford

Sample chromatograms of ambient air during the Abbotsford campaign are shown in Figs. 8 and 9. These particular chromatograms were chosen because they correspond to the periods with the highest (1.6 ppbv, Fig. 8) and lowest (8 pptv, Fig. 9) PAN mixing ratios. The latter was close to the PAN detection limit (~ 4 pptv during Abbotsford). In addition to the broad and off-scale O_2 peak eluting at ~ 0.5 min and the PAN and PPN peaks identified using standard injections, the chromatograms generally contained a large peak that eluted at ~ 1 min and another that eluted just after 3 min (a bit earlier to where APAN would have eluted). The areas of these peaks did not vary as much as those of the PANs, suggesting that the associated compounds are longer-lived

AMTD
7, 5953–6019, 2014

in the atmosphere than PAN (consistent with, for example, chlorofluorocarbons such as CCl_4).

When PAN concentrations were high, additional peaks appeared in the chromatograms. In Fig. 8, for example, the retention time of the peak eluting at 318 s is consistent with PiBN. Larger PANs, including MPAN, eluted after 6 min and appeared as ghost peaks in subsequent chromatograms (see, for example, the peaks around a retention time of 90 s in Fig. 8). Because their peak areas were small, they did not significantly interfere with the determination of PAN or PPN. However, the appearance of ghost peaks is undesirable and was eliminated by increasing the time between runs (from 6 to 10 min) and by the reducing the column length (from 20.41 m to 10.4 m) for the FOSSILS campaign (see Sect. 3.6 below).

Figure 10 shows the time series of PAN and PPN (top) and of O_3, NO and NO_2 (bottom) for the Abbotsford data set. PAN and PPN mixing ratios exhibited a strong diurnal profile that paralleled that of O_3, with maxima in the late afternoon. The PAN-GC and TD-CIMS data each cover the full dynamic range of the PAN mixing ratios in Abbotsford (from $\sim$ 8 pptv to 1.6 ppbv). For PPN, on the other hand, the lowest quantified mixing ratio from the PAN-GC was 16 pptv; hence, GC data are only available on days with relatively large PPN concentration. In contrast, the TD-CIMS was able to monitor PPN down to single pptv mixing ratios, and quantification of PPN by the PAN-GC was limited by the ability to fit the chromatographic peaks; for most of the Abbotsford data set, the PPN peak areas were too small to fit.

Figure 11 shows scatter plots of PAN-GC against TD-CIMS ambient air data. For PAN, the agreement between the PAN-GC and the raw TD-CIMS data is reasonable (r = 0.94, slope = 0.84 ± 0.01, Fig. 11a) and improves after the correction for TD-CIMS matrix effects arising from titration by NO and NO_2 is applied (r = 0.93, slope = 0.96 ± 0.01, Fig. 11c). For PPN, the uncorrected TD-CIMS data do not correlate well with PAN-GC data (r = 0.54, slope = 0.54 ± 0.05, Fig. 11b) but more so after the correction for inlet matrix effects is applied (r = 0.77, slope = 1.03 ± 0.07, Fig. 11d).

AMTD
7, 5953–6019, 2014

3.6 Measurement of PAN, PPN, and MPAN during FOSSILS

Figure 12 shows a typical chromatogram of ambient air and of a blank (i.e., when the PAN-GC sampled zero air) acquired during FOSSILS. The blank chromatogram demonstrates the absence of ghost peaks. The ambient air chromatogram shows peaks associated with PAN (121 pptv), PPN (13 pptv), and MPAN (10 pptv). The approximate elution time of APAN is noted, but its area was too small to be fitted. The retention times of peaks labeled with question marks do not correspond to any of the known PANs; they are likely either alkyl nitrates or halogenated compounds.

Figure 13 shows the time series of PAN, PPN, and MPAN mixing ratios observed during FOSSILS. In total, 3635 chromatograms were analyzed, with 3354 (2449, 2061) containing PAN (PPN, MPAN) at concentration levels above their respective detection limits (see below). The mixing ratios varied with time of day, generally peaked in the mid-afternoon and were highly correlated with each other, as expected. On average, the ratios of PPN/PAN and MPAN/PAN were $\sim 1:10$ and varied from $\sim 1:5$ to $\sim 1:20$. An analysis of these ratios to gain insights into the relative contributions of biogenic and anthropogenic hydrocarbons during the photochemical of O_3 (e.g., Roberts et al., 1998, 2002, 2007) will be presented elsewhere (manuscript in preparation).

At night, the PAN concentrations frequently dropped below quantifiable limits, because the relatively warm nocturnal temperatures facilitated thermal decomposition of PANs and because the high nocturnal NO concentrations from nearby anthropogenic activities titrated the peroxyacyl radicals. Data from such nights provided good opportunities to probe the PAN-GC detection limits. Table 12 summarizes fits of the PAN peak during the night of 3 September, which is an example of when PAN dropped near (and eventually below) quantifiable levels. Example fits are shown in Fig. 14. Analogous case studies for PPN and MPAN are shown in Figs. A4 and A5. The peaks in Fig. 14a and b are well above the limit of detection and quantification (even if the unsmoothed data had been used) and could have been fitted without any prior knowledge of the elution time and shape of the peak. In contrast, the fit of the peak shown in Fig. 14c

AMTD
7, 5953–6019, 2014

is less certain as the peak is barely above the background noise. In many cases, the fits converged only because the initial guess was set to the fit results of the preceding run (Table 12). On the basis of data as shown in Figs. 14, A4, and A5, the PAN-GC's limits of detection for PAN, PPN, and MPAN during FOSSILS were estimated at 1 pptv, 2 pptv, and 3 pptv, respectively.

The (slope) accuracy of the FOSSILS data is limited by the accuracy of the calibration, as the peak fitting procedure produces a (comparatively) small and hence negligible error. For PAN and PPN in particular, the limiting factor was the accuracy of the TD-CRDS, which is ±4 % (Paul and Osthoff, 2010). The accuracy of the MPAN and APAN data is conservatively estimated at ±15 % and ±25 %, respectively, because of the large uncertainties of calibration standard composition.

4 Discussion

4.1 Suitability of TD-CRDS to determine PAN-GC calibration factors

4.1.1 PAN and PPN

The linear correlation of PAN-GC peak areas with TD-CRDS PN mixing ratios over a wide concentration range (e.g., Fig. 7) affirms that TD-CRDS can be used as a straight-forward and easily implemented PAN-GC calibration method for PAN and PPN. The RF's obtained were reproducible over a period approaching one month which shows that both the TD-CRDS and PAN-GC responses remained stable over that time period.

Small differences in the relative RF's obtained for PAN and PPN were observed between the Abbotsford, FOSSILS, and post-FOSSILS laboratory work (Tables 3–5). This is to be expected because several parameters, for example the column oven temperature, sample flow rate, and analytical column length, differed between the campaigns. By-and-large, the results are consistent with each other and with those reported by

AMTD
7, 5953–6019, 2014

other groups in the literature (Table 7). For instance, the relative sensitivities of PPN to PAN during FOSSILS were $(0.91 \pm 0.02) : 1$ and agreed with the ratio of 0.90 ± 0.02 reported by Flocke et al. (2005); post-FOSSILS, the relative RF's of PPN : PAN were $(0.96 \pm 0.04) : 1$, slightly higher than during FOSSILS and slightly lower than in Abbotsford but still in agreement with Flocke et al. (2005).

Alkyl nitrates were also observed in the standard chromatograms. These molecules are formed as byproducts during the synthesis and are also decomposition products in aged standards. Because TD-CRDS can differentiate between PN and AN, the presence of alkyl nitrates in the standards could be tolerated, but would have constituted a significant interference if NO_y CL (without a preparatory-scale GC) had been used in the calibration. Likewise, NO_2, which may be present as a result of thermal decomposition of the PAN standard, is automatically quantified in the ambient temperature reference channel and subtracted (Paul et al., 2009; Paul and Osthoff, 2010).

4.1.2 MPAN

The challenges with calibrating a PAN-GC for MPAN (or APAN) are well-documented in the literature (e.g., Flocke et al., 2005; Zheng et al., 2011; Roberts et al., 2001) and are associated mainly with the difficult delivery of a high-purity gaseous sample. The use of the TD-CRDS calibration method has the inherent advantage that the 250 °C channel is insensitive to many of the impurities generated during the MPAN (or APAN) synthesis, i.e., alkyl nitrates, nitryl chloride, and nitric acid (Paul et al., 2009; Paul and Osthoff, 2010). Thus, the TD-CRDS calibration method only requires quantification of (and correction for) PAN impurities.

To determine the RF for MPAN, the PN data had to be corrected for PAN and PPN, which were quantified using the APAN standard chromatogram directly (and their RF's determined earlier). This process was somewhat laborious but otherwise straightforward. The presence of PAN in synthetic MPAN standards that was observed in all MPAN standards analyzed in this work has been observed before by several groups (Grosjean et al., 1993a; Flocke et al., 2005; Bertman and Roberts, 1991). Flocke

AMTD
7, 5953–6019, 2014

et al. (2005) suggested that PAN is produced during the synthesis from impurities in the reagents (i.e., acetic anhydride). While it is possible that PPN is produced in a similar fashion during the synthesis (i.e., from propanoic anhydride which would be present as a trace impurity), PPN originated mostly from memory effects within the Teflon tubing in this work.

Having corrected the MPAN data for PAN and PPN, the MPAN RF (relative to PAN) was determined at $(0.83 \pm 0.04) : 1$ and $(0.80 \pm 0.02) : 1$ (Tables 4 and 5). These values are $\sim 25\,\%$ greater than the ratio of $(0.64 \pm 0.03) : 1$ reported by Flocke et al. (2005) and somewhat greater than the $0.72 : 1$ ratio reported by Williams et al. (2000). The reason for the discrepancy between their and our RF is not clear. It is unlikely that a systematic error from the subtraction of PAN (or PPN) biased the results high since the RF for MPAN stayed within a narrow range throughout the FOSSILS campaign, even though the PAN content in the MPAN standard varied from 73 % at the beginning to 6 % at the end of the campaign (concentrations of PPN were generally too small to affect the slope of the calibration plots). It is therefore likely that there are real differences in the sensitivity between the instruments, e.g., due to differences in on-column decomposition rates (Roumelis and Glavas, 1989) between the RTX-200 column used by the other groups and the RTX-1701 column used in this work.

Overall, we conclude on the basis of the linear correlation of PN mixing ratios with PAN-GC peak areas that TD-CRDS can be used to calibrate a PAN-GC for MPAN, as was demonstrated in this work with a much greater uncertainty (±15 %) than the TD-CRDS would allow for in principle (±4 %). In the future, we intend to construct a preparatory-scale GC similar to the one developed at NOAA and NCAR (Williams et al., 2000) to clean up the MPAN output used in the calibration to reduce the uncertainty of the calibration and to eliminate the need to subtract PAN impurities from the PN mixing ratio.

AMTD
7, 5953–6019, 2014

4.1.3 APAN

For APAN, only a lower limit to the PAN-GC RF could be determined in this work because the PN measurement was compromised by unusual PAN-like compounds co-emitted by the diffusion source (Fig. 6), many of which (e.g., the chlorinated adduct) could have formed during the synthesis. Their presence is corroborated by multiple impurity peaks observed in the APAN standard chromatogram (Fig. 5a). However, in TD-CIMS, counts at certain masses can be positively biased by proton exchange reactions of carboxylate product ions with acid impurities, and these signals can masquerade as PANs (Mielke and Osthoff, 2012) such that the impurities in Fig. 6 cannot be quantified with any certainty. If these impurities are taken into account, the RF relative to PAN is estimated to be between 0.85 and 1. This range brackets the RF of 0.959 ± 0.006 reported by Zheng et al. (2011) for their PAN-GC. This consistency and the high linearity of the calibration plots suggest that TD-CRDS can, in principle, be used for accurate APAN calibrations (even though it wasn't demonstrated in this work). The use of the aforementioned preparatory GC in future experiment will hopefully confirm this.

APAN was synthesized from its acyl chloride in this study. Apparently, $ClNO_2$ is formed during the synthesis from reaction between the chloride anion (Cl^-) liberated from acryloyl chloride and the concentrated sulfuric/nitric acid mixture which generates the nitronium cation, NO_2^+ (Behnke et al., 1997), and $ClNO_2$ is sufficiently nonpolar to be extracted from the aqueous solution into tridecane during the workup. Evaporation of $ClNO_2$ from a standard would interfere with a calibration based on NO_y chemiluminescence, but does not in a TD-CRDS PN measurement (Thaler et al., 2011). Furthermore, $ClNO_2$ autoionizes to Cl^- and NO_2^+ (Behnke et al., 1997) which can add across C-C double bonds and, for example, accelerate the degradation of APAN in the standard solution. The presence of $ClNO_2$ in PAN standards is undesirable and may potentially explain the difficulties documented by, for example, Zheng et al. (2011), and encountered in this work with the determination of APAN response factors. We conclude that PAN syntheses that use acyl chlorides as starting material should therefore

AMTD
7, 5953–6019, 2014

be avoided. The MPAN standard used by Flocke et al. (2005) and Zheng et al. (2011) was synthesized from methacryloyl chloride and likely also contained $ClNO_2$; it is unclear, however, if (and when) $ClNO_2$ would elute from their preparatory GC column and if the results were affected by this impurity.

4.2 Performance of the PAN-GC in the field

4.2.1 Chromatography

The majority of PAN-GC in use today utilizes a RTX-200 analytical column (or similar). Tanimoto and coworkers (Tanimoto et al., 1999; Tanimoto and Akimoto, 2001) and Flocke et al. (2005) have described PAN instruments with RTX-1701 columns. While the absence of co-eluting peaks was demonstrated for RTX-200 columns (Williams et al., 2000), no explicit mention has been made so far in the literature confirming the absence of co-eluting peaks (which would positively interfere) for RTX-1701 columns. The agreement of PAN-GC and TD-CIMS data corroborate the (assumed) absence of significant interferences. Nevertheless, a PAN inlet scrubber as shown by (Williams et al., 2000) will be implemented for future deployments as a precaution.

There are many PAN-GC's that are equipped with guard columns (e.g., Pätz et al., 2002; Flocke et al., 2005). The absence of any degradation of performance suggests guard columns are not needed, unless one was to perhaps sample in an extraordinarily polluted environment.

The sample chromatograms (Figs. 5, 8, 9, and 12) show that PANs are sufficiently resolved such that a ~ 50 % shorter column could have been used to speed up the analysis and increase the peak height (and hence improve time resolution and lower the detection limit). A shorter column will be deployed in future studies.

AMTD
7, 5953–6019, 2014

4.2.2 Peak fitting procedure

Conventionally, quantification in GC is performed using peak height or by integration of the peak area. The fitting procedure used in this work deviates from this approach as the parameters (one of which is the peak area, *A*) of the Gaussian expression Eq. (1) are fitted to the data, a procedure that yields results that are quite precise. One has to keep in mind that, unlike for linear fits, there are no analytical solutions to describe uncertainties in multivariate fits, and that the errors reported by the software package Igor Pro are error estimates only. We suspect the error introduced by the fitting procedure if, for example, analyzed by Monte-Carlo techniques, is somewhat larger than stated but in all likelihood less than the uncertainties of the TD-CRDS and arising from standard impurities.

4.2.3 Limits of detection and accuracy of measurements

The limits of detection of the University of Calgary PAN-GC are in the single digit pptv range and of similar magnitude as, or marginally better than, those reported for other PAN-GC instruments that have recently been operated in the field (e.g., Thornberry et al., 2001; Roberts et al., 2002, 2006, Table 13). Unfortunately, there is no consensus in the literature on how best to determine LODs in gas chromatography, and methods such as signal-to-noise, blank determination, and linear regression methods often give different results (e.g., Sanagi et al., 2009). In this work, the LODs were estimated from the smallest peak area that could be fitted (Figs. 14, A4 and A5), which is a somewhat subjective evaluation. A more rigorous determination would have required repeated measurements near the LOD; however, for this procedure to succeed, one would have to assume that the output of a low-concentration standard is sufficiently stable to not add to the measurement noise, which is difficult to demonstrate.

The accuracy of the FOSSILS data is limited by the accuracy of the calibration, as the fitting procedure produces a (relatively) negligible error. For PAN and PPN, the limiting factor is the accuracy of the TD-CRDS, which is ±4 % (Paul and Osthoff, 2010),

AMTD
7, 5953–6019, 2014

so that the accuracy of the PAN and PPN data is ±4 % ±1 pptv. For MPAN and APAN, the accuracy is conservatively estimated at ±15 % ±2 pptv and ±25 % ±3 pptv, driven mainly by the large uncertainties of calibration standard composition. If these were to be improved, the accuracy of the MPAN and APAN PAN-GC measurements could likely be improved further (e.g., to ±5 %).

Stability of the sample flow rate was identified as a factor affecting the RF in the PAN-GC. When the PAN-GC was operated with a slower sample flow rate (Table 5, Fig. 7b), the RF's were systematically greater than those when the PAN-GC was operated with a faster sample flow rate (Table 4, Fig. 7a) which lowered the number of molecules injected onto the column. In addition, the standard widths of the fits of the PAN peak were ~ 20 % larger (Fig. A1) which suggests that the greater amount of air sampled led to a small amount of extra column band broadening.

4.3 Comparison of PAN-GC with TD-CIMS data

Zheng et al. (2011) and Mielke and Osthoff (2012) have documented some of the analytical challenges of using TD-CIMS to make accurate measurements of the various PANs. In Abbotsford, the TD-CIMS data was compromised by the lack of bona fide ^{13}C-labeled PAN internal standard but nevertheless showed excellent agreement with the PAN-GC's PAN and PPN data with slopes close to unity (Figs. 10 and 11). Part of the scatter may be due to temporal mismatches in the timing of the measurements. The general good agreement suggests that interference from acids as encountered in Pasadena (Mielke and Osthoff, 2012) played at best a minor role in the Abbotsford data set. Interference from peroxy acids (Phillips et al., 2013) was accounted for by subtracting the background measured in high NO concentration.

The limits of detection for PAN and PPN for TD-CIMS and PAN-GC (after the modifications for the FOSSILS campaign) are both in the single digit pptv range, with the TD-CIMS having much faster time response (Slusher et al., 2004; Turnipseed et al., 2006) which is an advantage when sampling rapidly changing air masses, for example, encountered from a moving platform such as an aircraft. For MPAN, however,

AMTD
7, 5953–6019, 2014

PAN-GCs generally have the advantage because TD-CIMS struggles with quantification of MPAN in ambient air due to an interference from CPAN (Zheng et al., 2011) and a poor temperature-dependent response factor at $m\ z$ 85 (Table 9).

5 Summary and conclusions

A state-of-the-art PAN-GC has been described and calibrated using TD-CRDS for PAN, PPN MPAN, and APAN. The TD-CRDS method is easily implemented and more tolerant to impurities than the conventionally used NO_y CL method, but is not entirely immune to PAN impurities co-generated during synthesis; hence, a preparatory GC will be constructed for future calibrations. Synthetic routes with acyl chlorides as starting reagents were found to generate $ClNO_2$ and should not be used to generate PAN standards. PAN-GC is preferred over TD-CIMS in the quantification of MPAN in ambient air.

Acknowledgements. This project was undertaken with the financial support of the Government of Canada through the Federal Department of the Environment. Ce projet a été réalisé avec l'appui financier du Gouvernement du Canada agissant par l'entremise du ministère fédéral de l'Environnement. Partial funding for this work was provided by the Natural Sciences and Engineering Research Council of Canada (NSERC) in the form of operating ("Discovery") and Research Tools and Instruments (RTI) grants. The Abbotsford field study was financially supported by a research grant from the Fraser Basin Council of British Columbia, and by Metro Vancouver. The authors thank J. Love and C. Berlinguette for a GC-ECD that was converted to a PAN-GC for this work.

AMTD
7, 5953–6019, 2014

References

Ainslie, B. and Steyn, D. G.: Spatiotemporal trends in episodic ozone pollution in the lower Fraser Valley, British Columbia, in relation to mesoscale atmospheric circulation patterns and emissions, J. Appl. Meteorol. Clim., 46, 1631–1644, doi:10.1175/JAM2547.1, 2007.

Altshuller, A. P.: PANs in the atmosphere, J. Air Waste Manag. Assoc., 43, 1221–1230, doi:10.1080/1073161X.1993.10467199, 1993.

Alvarado, M. J., Logan, J. A., Mao, J., Apel, E., Riemer, D., Blake, D., Cohen, R. C., Min, K.-E., Perring, A. E., Browne, E. C., Wooldridge, P. J., Diskin, G. S., Sachse, G. W., Fuelberg, H., Sessions, W. R., Harrigan, D. L., Huey, G., Liao, J., Case-Hanks, A., Jimenez, J. L., Cubison, M. J., Vay, S. A., Weinheimer, A. J., Knapp, D. J., Montzka, D. D., Flocke, F. M., Pollack, I. B., Wennberg, P. O., Kurten, A., Crounse, J., Clair, J. M. St., Wisthaler, A., Mikoviny, T., Yantosca, R. M., Carouge, C. C., and Le Sager, P.: Nitrogen oxides and PAN in plumes from boreal fires during ARCTAS-B and their impact on ozone: an integrated analysis of aircraft and satellite observations, Atmos. Chem. Phys., 10, 9739–9760, doi:10.5194/acp-10-9739-2010, 2010.

Behnke, W., George, C., Scheer, V., and Zetzsch, C.: Production and decay of $ClNO_2$, from the reaction of gaseous N_2O_5 with NaCl solution: bulk and aerosol experiments, J. Geophys. Res., 102, 3795–3804, doi:10.1029/96JD03057, 1997.

Bertman, S. B. and Roberts, J. M.: A PAN analog from isoprene photooxidation, Geophys. Res. Lett., 18, 1461–1464, doi:10.1029/91gl01852, 1991.

Blanchard, P., Shepson, P. B., So, K. W., Schiff, H. I., Bottenheim, J. W., Gallant, A. J., Drummond, J. W., and Wong, P.: A comparison of calibration and measurement techniques for gas chromatographic determination of atmospheric peroxyacetyl nitrate (PAN), Atmos. Environ., 24, 2839–2846, doi:10.1016/0960-1686(90)90171-I, 1990.

Darley, E. F., Kettner, K. A., and Stephens, E. R.: Analysis of Peroxyacyl Nitrates by Gas Chromatography with Electron Capture Detection, Anal. Chem., 35, 589–591, doi:10.1021/ac60197a028, 1963.

Day, D. A., Wooldridge, P. J., Dillon, M. B., Thornton, J. A., and Cohen, R. C.: A thermal dissociation laser-induced fluorescence instrument for in situ detection of NO_2, peroxy nitrates, alkyl nitrates, and HNO_3, J. Geophys. Res., 107, 4046, doi:10.1029/2001JD000779, 2002.

Eisele, F. L., Mauldin, L., Cantrell, C., Zondlo, M., Apel, E., Fried, A., Walega, J., Shetter, R., Lefer, B., Flocke, F., Weinheimer, A., Avery, M., Vay, S., Sachse, G., Podolske, J., Diskin, G.,

AMTD
7, 5953–6019, 2014

Barrick, J. D., Singh, H. B., Brune, W., Harder, H., Martinez, M., Bandy, A., Thornton, D., Heikes, B., Kondo, Y., Riemer, D., Sandholm, S., Tan, D., Talbot, R., and Dibb, J.: Summary of measurement intercomparisons during TRACE-P, J. Geophys. Res.-Atmos., 108, 8791, doi:10.1029/2002jd003167, 2003.

Emrich, M. and Warneck, P.: Photodissociation of acetone in air: dependence on pressure and wavelength. Behavior of the excited singlet state, J. Phys. Chem. A, 104, 9436–9442, doi:10.1021/jp001873i, 2000.

Everest, M. A. and Atkinson, D. B.: Discrete sums for the rapid determination of exponential decay constants, Rev. Sci. Instrum., 79, 023108–023109, doi:10.1063/1.2839918, 2008.

Fischer, E. V., Jaffe, D. A., Reidmiller, D. R., and Jaegle, L.: Meteorological controls on observed peroxyacetyl nitrate at Mount Bachelor during the spring of 2008, J. Geophys. Res., 115, D03302, doi:10.1029/2009JD012776, 2010.

Flocke, F. M., Weinheimer, A. J., Swanson, A. L., Roberts, J. M., Schmitt, R., and Shertz, S.: On the measurement of PANs by gas chromatography and electron capture detection, J. Atmos. Chem., 52, 19–43, doi:10.1007/s10874-005-6772-0, 2005.

Furgeson, A., Mielke, L. H., Paul, D., and Osthoff, H. D.: A photochemical source of peroxypropionic and peroxyisobutanoic nitric anhydride, Atmos. Environ., 45, 5025–5032, doi:10.1016/j.atmosenv.2011.03.072, 2011.

Gaffney, J. S., Fajer, R., and Senum, G. I.: An improved procedure for high-purity gaseous peroxyacyl nitrate production – use of heavy lipid solvents, Atmos. Environ., 18, 215–218, doi:10.1016/0004-6981(84)90245-2, 1984.

Gregory, G. L., Hoell, J. M., Ridley, B. A., Singh, H. B., Gandrud, B., Salas, L. J., and Shetter, J.: An intercomparison of airborne PAN measurements, J. Geophys. Res.-Atmos., 95, 10077–10087, doi:10.1029/JD095iD07p10077, 1990.

Grosjean, D., Fung, K., Collins, J., Harrison, J., and Breitung, E.: Portable generator for on-site calibration of peroxyacetyl nitrate analyzers, Anal. Chem., 56, 569–573, doi:10.1021/ac00267a059, 1984.

Grosjean, D., Williams, E. L., and Grosjean, E.: Atmospheric chemistry of isoprene and of its carbonyl products, Environ. Sci. Technol., 27, 830–840, doi:10.1021/es00042a004, 1993a.

Grosjean, D., Williams, E. L., and Grosjean, E.: Peroxyacyl nitrates at southern California mountain forest locations, Environ. Sci. Technol., 27, 110–121, doi:10.1021/es00038a011, 1993b.

AMTD
7, 5953–6019, 2014

Grosjean, D., Grosjean, E., and Williams, E. L.: Thermal decomposition of PAN, PPN and vinyl-PAN, J. Air Waste Manag. Assoc., 44, 391–396, doi:10.1080/1073161X.1994.10467260, 1994.

Hansel, A. and Wisthaler, A.: A method for real-time detection of PAN, PPN and MPAN in ambient air, Geophys. Res. Lett., 27, 895–898, doi:10.1029/1999GL010989, 2000.

Hastie, D. R., Gray, J., Langford, V. S., Maclagan, R., Milligan, D. B., and McEwan, M. J.: Real-time measurement of peroxyacetyl nitrate using selected ion flow tube mass spectrometry, Rapid Commun. Mass Sp., 24, 343–348, doi:10.1002/rcm.4400, 2010.

Horvath, S. M., Bedi, J. F., and Drechslerparks, D. M.: Effects of peroxyacetyl nitrate alone and in combination with ozone in healthy-young women, J. Air Pollut. Control Assoc., 36, 265–270, doi:10.1080/00022470.1986.10466067, 1986.

Kirchner, F., Mayer-Figge, A., Zabel, F., and Becker, K. H.: Thermal stability of peroxynitrates, Int. J. Chem. Kin., 31, 127–144, 1999.

Kleindienst, T. E.: Recent developments in the chemistry and biology of peroxyacetyl nitrate, Res. Chem. Intermed., 20, 335–384, doi:10.1163/156856794X00379, 1994.

LaFranchi, B. W., Wolfe, G. M., Thornton, J. A., Harrold, S. A., Browne, E. C., Min, K. E., Wooldridge, P. J., Gilman, J. B., Kuster, W. C., Goldan, P. D., de Gouw, J. A., McKay, M., Goldstein, A. H., Ren, X., Mao, J., and Cohen, R. C.: Closing the peroxy acetyl nitrate budget: observations of acyl peroxy nitrates (PAN, PPN, and MPAN) during BEARPEX 2007, Atmos. Chem. Phys., 9, 7623–7641, doi:10.5194/acp-9-7623-2009, 2009.

Marley, N. A., Gaffney, J. S., White, R. V., Rodriguez-Cuadra, L., Herndon, S. E., Dunlea, E., Volkamer, R. M., Molina, L. T., and Molina, M. J.: Fast gas chromatography with luminol chemiluminescence detection for the simultaneous determination of nitrogen dioxide and peroxyacetyl nitrate in the atmosphere, Rev. Sci. Instrum., 75, 4595–4605, doi:10.1063/1.1805271, 2004.

McNeill, V. F., Patterson, J., Wolfe, G. M., and Thornton, J. A.: The effect of varying levels of surfactant on the reactive uptake of N_2O_5 to aqueous aerosol, Atmos. Chem. Phys., 6, 1635–1644, doi:10.5194/acp-6-1635-2006, 2006.

Mielke, L. H. and Osthoff, H. D.: On quantitative measurements of peroxycarboxylic nitric anhydride mixing ratios by thermal dissociation chemical ionization mass spectrometry, Int. J. Mass Spectrom., 310, 1–9, doi:10.1016/j.ijms.2011.10.005, 2012.

Mielke, L. H., Furgeson, A., and Osthoff, H. D.: Observation of $ClNO_2$ in a mid-continental urban environment, Environ. Sci. Technol., 45, 8889–8896, doi:10.1021/es201955u, 2011.

AMTD
7, 5953–6019, 2014

Mielke, L. H., Stutz, J., Tsai, C., Hurlock, S. C., Roberts, J. M., Veres, P. R., Froyd, K. D., Hayes, P. L., Cubison, M. J., Jimenez, J. L., Washenfelder, R. A., Young, C. J., Gilman, J. B., de Gouw, J. A., Flynn, J. H., Grossberg, N., Lefer, B. L., Liu, J., Weber, R. J., and Osthoff, H. D.: Heterogeneous formation of nitryl chloride and its role as a nocturnal NO_x reservoir species during CalNex-LA 2010, J. Geophys. Res., 118, 10638–10652, doi:10.1002/jgrd.50783, 2013.

Mills, G. P., Sturges, W. T., Salmon, R. A., Bauguitte, S. J.-B., Read, K. A., and Bandy, B. J.: Seasonal variation of peroxyacetylnitrate (PAN) in coastal Antarctica measured with a new instrument for the detection of sub-part per trillion mixing ratios of PAN, Atmos. Chem. Phys., 7, 4589–4599, doi:10.5194/acp-7-4589-2007, 2007.

Nikitas, C., Clemitshaw, K. C., Oram, D. E., and Penkett, S. A.: Measurements of PAN in the polluted boundary layer and free troposphere using a luminol-NO_2 detector combined with a thermal converter, J. Atmos. Chem., 28, 339–359, doi:10.1023/a:1005898017520, 1997.

Nouaime, G., Bertman, S. B., Seaver, C., Elyea, D., Huang, H., Shepson, P. B., Starn, T. K., Riemer, D. D., Zika, R. G., and Olszyna, K.: Sequential oxidation products from tropospheric isoprene chemistry: MACR and MPAN at a NO_x-rich forest environment in the southeastern United States, J. Geophys. Res., 103, 22463–22471, doi:10.1029/98JD00320, 1998.

Odame-Ankrah, C. A. and Osthoff, H. D.: A compact diode laser cavity ring-down spectrometer for atmospheric measurements of NO_3 and N_2O_5 with automated zeroing and calibration, Appl. Spectrosc., 65, 1260–1268, doi:10.1366/11-06384, 2011.

Osthoff, H. D., Roberts, J. M., Ravishankara, A. R., Williams, E. J., Lerner, B. M., Sommariva, R., Bates, T. S., Coffman, D., Quinn, P. K., Stark, H., Burkholder, J. B., Talukdar, R. K., Meagher, J., Fehsenfeld, F. C., and Brown, S. S.: High levels of nitryl chloride in the polluted subtropical marine boundary layer, Nat. Geosci., 1, 324–328, doi:10.1038/ngeo177, 2008.

Pandey Deolal, S., Staehelin, J., Brunner, D., Cui, J., Steinbacher, M., Zellweger, C., Henne, S., and Vollmer, M. K.: Transport of PAN and NOy from different source regions to the Swiss high alpine site Jungfraujoch, Atmos. Environ., 64, 103–115, doi:10.1016/j.atmosenv.2012.08.021, 2013.

Pätz, H.-W., Lerner, A., Houben, N., and Volz-Thomas, A.: Validation of a new method for the calibration of peroxy acetyl nitrate (PAN)-analyzers, Gefahrst. Reinhalt. L., 62, 215–219, 2002.

AMTD
7, 5953–6019, 2014

Paul, D. and Osthoff, H. D.: Absolute Measurements of Total Peroxy Nitrate Mixing Ratios by Thermal Dissociation Blue Diode Laser Cavity Ring-Down Spectroscopy, Anal. Chem., 82, 6695–6703, doi:10.1021/ac101441z, 2010.

Paul, D., Furgeson, A., and Osthoff, H. D.: Measurement of total alkyl and peroxy nitrates by thermal decomposition cavity ring-down spectroscopy, Rev. Sci. Instrum., 80, 114101, doi:10.1063/1.3258204, 2009.

Phillips, G. J., Pouvesle, N., Thieser, J., Schuster, G., Axinte, R., Fischer, H., Williams, J., Lelieveld, J., and Crowley, J. N.: Peroxyacetyl nitrate (PAN) and peroxyacetic acid (PAA) measurements by iodide chemical ionisation mass spectrometry: first analysis of results in the boreal forest and implications for the measurement of PAN fluxes, Atmos. Chem. Phys., 13, 1129–1139, doi:10.5194/acp-13-1129-2013, 2013.

Roberts, J. M.: The atmospheric chemistry of organic nitrates, Atmos. Environ., 24, 243–287, doi:10.1016/0960-1686(90)90108-Y, 1990.

Roberts, J. M.: PAN and Related Compounds, in: Volatile Organic Compounds in the Atmosphere, edited by: Koppmann, R., Blackwell Publishing, Oxford, UK, 221–268, 2007.

Roberts, J. M., Williams, J., Baumann, K., Buhr, M. P., Goldan, P. D., Holloway, J., Hubler, G., Kuster, W. C., McKeen, S. A., Ryerson, T. B., Trainer, M., Williams, E. J., Fehsenfeld, F. C., Bertman, S. B., Nouaime, G., Seaver, C., Grodzinsky, G., Rodgers, M., and Young, V. L.: Measurements of PAN, PPN, and MPAN made during the 1994 and 1995 Nashville Intensives of the Southern Oxidant Study: implications for regional ozone production from biogenic hydrocarbons, J. Geophys. Res., 103, 22473–22490, doi:10.1029/98JD01637, 1998.

Roberts, J. M., Flocke, F., Weinheimer, A., Tanimoto, H., Jobson, B. J., Riemer, D., Apel, E., Atlas, E., Donnelly, S., Stroud, V., Johnson, K., Weaver, R., and Fehsenfeld, F. C.: Observations of APAN during TexAQS 2000, Geophys. Res. Lett., 28, 4195–4198, doi:10.1029/2001GL013466, 2001.

Roberts, J. M., Flocke, F., Stroud, C. A., Hereid, D., Williams, E., Fehsenfeld, F., Brune, W., Martinez, M., and Harder, H.: Ground-based measurements of peroxycarboxylic nitric anhydrides (PANs) during the 1999 Southern Oxidants Study Nashville Intensive, J. Geophys. Res., 107, 4554, doi:10.1029/2001jd000947, 2002.

Roberts, J. M., Jobson, B. T., Kuster, W., Goldan, P., Murphy, P., Williams, E., Frost, G., Riemer, D., Apel, E., Stroud, C., Wiedinmyer, C., and Fehsenfeld, F.: An examination of the chemistry of peroxycarboxylic nitric anhydrides and related volatile organic compounds

AMTD
7, 5953–6019, 2014

during Texas Air Quality Study 2000 using ground-based measurements, J. Geophys. Res.-Atmos., 108, 4495, doi:10.1029/2003jd003383, 2003.
Roberts, J. M., Marchewka, M., Bertman, S. B., Goldan, P., Kuster, W., de Gouw, J., Warneke, C., Williams, E., Lerner, B., Murphy, P., Apel, E., and Fehsenfeld, F. C.: Analysis of the isoprene chemistry observed during the New England Air Quality Study (NEAQS) 2002 intensive experiment, J. Geophys. Res., 111, D23S12, doi:10.1029/2006JD007570, 2006.
Roberts, J. M., Marchewka, M., Bertman, S. B., Sommariva, R., Warneke, C., de Gouw, J., Kuster, W., Goldan, P., Williams, E., Lerner, B. M., Murphy, P., and Fehsenfeld, F. C.: Measurements of PANs during the New England Air Quality Study 2002, J. Geophys. Res.-Atmos., 112, D20306, doi:10.1029/2007jd008667, 2007.
Roumelis, N. and Glavas, S.: Decomposition of peroxyacetyl nitrate and peroxypropionyl nitrate during gas chromatographic determination with a wide-bore capillary and two packed columns, Anal. Chem., 61, 2731–2734, doi:10.1021/ac00199a010, 1989.
Sanagi, M. M., Ling, S. L., Nasir, Z., Hermawan, D., Ibrahim, W. A. W., and Abu Naim, A.: Comparison of Signal-to-Noise, Blank Determination, and Linear Regression Methods for the Estimation of Detection and Quantification Limits for Volatile Organic Compounds by Gas Chromatography, J. AOAC Int., 92, 1833–1838, 2009.
Schrimpf, W., Müller, K. P., Johnen, F. J., Lienaerts, K., and Rudolph, J.: An optimized method for airborne peroxyacetyl nitrate (PAN) measurements, J. Atmos. Chem., 22, 303–317, doi:10.1007/BF00696640, 1995.
Slusher, D. L., Huey, L. G., Tanner, D. J., Flocke, F. M., and Roberts, J. M.: A thermal dissociation-chemical ionization mass spectrometry (TD-CIMS) technique for the simultaneous measurement of peroxyacyl nitrates and dinitrogen pentoxide, J. Geophys. Res., 109, D19315, doi:10.1029/2004JD004670, 2004.
Stroud, C., Madronich, S., Atlas, E., Ridley, B., Flocke, F., Weinheimer, A., Talbot, B., Fried, A., Wert, B., Shetter, R., Lefer, B., Coffey, M., Heikes, B., and Blake, D.: Photochemistry in the arctic free troposphere: NOx budget and the role of odd nitrogen reservoir recycling, Atmos. Environ., 37, 3351–3364, doi:10.1016/S1352-2310(03)00353-4, 2003.
Taha, Y. M., Odame-Ankrah, C. A., and Osthoff, H. D.: Real-time vapor detection of nitroaromatic explosives by catalytic thermal dissociation blue diode laser cavity ring-down spectroscopy, Chem. Phys. Lett., 582, 15–20, doi:10.1016/j.cplett.2013.07.040, 2013.

AMTD
7, 5953–6019, 2014

Tanimoto, H. and Akimoto, H.: A new peroxycarboxylic nitric anhydride identified in the atmosphere: CH_2=CHC(O)OONO_2 (APAN), Geophys. Res. Lett., 28, 2831–2834, doi:10.1029/2001GL012998, 2001.

Tanimoto, H., Hirokawa, J., Kajii, Y., and Akimoto, H.: A new measurement technique of peroxyacetyl nitrate at parts per trillion by volume levels: gas chromatography/negative ion chemical ionization mass spectrometry, J. Geophys. Res., 104, 21343–21354, doi:10.1029/1999JD900345, 1999.

Thaler, R. D., Mielke, L. H., and Osthoff, H. D.: Quantification of nitryl chloride at part per trillion mixing ratios by thermal dissociation cavity ring-down spectroscopy, Anal. Chem., 83, 2761–2766, doi:10.1021/ac200055z, 2011.

Thornberry, T., Carroll, M. A., Keeler, G. J., Sillman, S., Bertman, S. B., Pippin, M. R., Ostling, K., Grossenbacher, J. W., Shepson, P. B., Cooper, O. R., Moody, J. L., and Stockwell, W. R.: Observations of reactive oxidized nitrogen and speciation of NO_y during the PROPHET summer 1998 intensive, J. Geophys. Res., 106, 24359–24386, doi:10.1029/2000JD900760, 2001.

Turnipseed, A. A., Huey, L. G., Nemitz, E., Stickel, R., Higgs, J., Tanner, D. J., Slusher, D. L., Sparks, J. P., Flocke, F., and Guenther, A.: Eddy covariance fluxes of peroxyacetyl nitrates (PANs) and NO_y to a coniferous forest, J. Geophys. Res., 111, D09304, doi:10.1029/2005JD006631, 2006.

Vingarzan, R. and Li, S. M.: The Pacific 2001 air quality study – synthesis of findings and policy implications, Atmos. Environ., 40, 2637–2649, doi:10.1016/j.atmosenv.2005.09.083, 2006.

Volz-Thomas, A., Xueref, I., and Schmitt, R.: An automatic gas chromatograph and calibration system for ambient measurements of PAN and PPN, Environ. Sci. Poll. Res., 9, 72–76, 2002.

Warneck, P. and Zerbach, T.: Synthesis of peroxyacetyl nitrate in air by acetone photolysis, Environ. Sci. Technol., 26, 74–79, doi:10.1021/es00025a005, 1992.

Williams, E. L., Grosjean, E., and Grosjean, D.: Ambient levels of the peroxyacyl nitrates PAN, PPN and MPAN in Atlanta, Georgia, J. Air Waste Manag. Assoc., 43, 873–879, doi:10.1080/1073161X.1993.10467170, 1993.

Williams, J., Roberts, J. M., Bertman, S. B., Stroud, C. A., Fehsenfeld, F. C., Baumann, K., Buhr, M. P., Knapp, K., Murphy, P. C., Nowick, M., and Williams, E. J.: A method for the airborne measurement of PAN, PPN, and MPAN, J. Geophys. Res., 105, 28943–28960, doi:10.1029/2000JD900373, 2000.

Wolfe, G. M., Thornton, J. A., McNeill, V. F., Jaffe, D. A., Reidmiller, D., Chand, D., Smith, J., Swartzendruber, P., Flocke, F., and Zheng, W.: Influence of trans-Pacific pollution transport

AMTD
7, 5953–6019, 2014

on acyl peroxy nitrate abundances and speciation at Mount Bachelor Observatory during INTEX-B, Atmos. Chem. Phys., 7, 5309–5325, doi:10.5194/acp-7-5309-2007, 2007.

Wolfe, G. M., Thornton, J. A., Yatavelli, R. L. N., McKay, M., Goldstein, A. H., LaFranchi, B., Min, K.-E., and Cohen, R. C.: Eddy covariance fluxes of acyl peroxy nitrates (PAN, PPN and MPAN) above a Ponderosa pine forest, Atmos. Chem. Phys., 9, 615–634, doi:10.5194/acp-9-615-2009, 2009.

Wooldridge, P. J., Perring, A. E., Bertram, T. H., Flocke, F. M., Roberts, J. M., Singh, H. B., Huey, L. G., Thornton, J. A., Wolfe, G. M., Murphy, J. G., Fry, J. L., Rollins, A. W., LaFranchi, B. W., and Cohen, R. C.: Total Peroxy Nitrates (PNs) in the atmosphere: the Thermal Dissociation-Laser Induced Fluorescence (TD-LIF) technique and comparisons to speciated PAN measurements, Atmos. Meas. Tech., 3, 593–607, doi:10.5194/amt-3-593-2010, 2010.

Zhang, G., Mu, Y. J., Liu, J. F., and Mellouki, A.: Direct and simultaneous determination of trace-level carbon tetrachloride, peroxyacetyl nitrate, and peroxypropionyl nitrate using gas chromatography-electron capture detection, J. Chromatogr. A, 1266, 110–115, doi:10.1016/j.chroma.2012.09.092, 2012.

Zhang, J. M., Wang, T., Ding, A. J., Zhou, X. H., Xue, L. K., Poon, C. N., Wu, W. S., Gao, J., Zuo, H. C., Chen, J. M., Zhang, X. C., and Fan, S. J.: Continuous measurement of peroxyacetyl nitrate (PAN) in suburban and remote areas of western China, Atmos. Environ., 43, 228–237, doi:10.1016/j.atmosenv.2008.09.070, 2009.

Zheng, W., Flocke, F. M., Tyndall, G. S., Swanson, A., Orlando, J. J., Roberts, J. M., Huey, L. G., and Tanner, D. J.: Characterization of a thermal decomposition chemical ionization mass spectrometer for the measurement of peroxy acyl nitrates (PANs) in the atmosphere, Atmos. Chem. Phys., 11, 6529–6547, doi:10.5194/acp-11-6529-2011, 2011.

AMTD
7, 5953–6019, 2014

Table 1. Parameters of Eq. (1) fitted to the top 3 chromatograms in Fig. 5a. The fit uncertainties are those estimated using built-in algorithms in the software package Igor Pro 6. The uncertainty stated for PN is the 1σ measurement precision. Concentrations of impurities were not determined.

Parameter (unit)	PAN	APAN	PPN
t_0 (s)	76.379 ± 0.003	133.570 ± 0.002	160.75 ± 0.010
σ (s)	1.238 ± 0.004	1.882 ± 0.003	2.15 ± 0.012
V_0 (mV)*	250.5 ± 1.4	7.3 ± 4.0	164.4 ± 1.8
m (mV s^{-1})	0.017 ± 0.017	0.72 ± 0.03	0.105 ± 0.011
A (mV s)	350.6 ± 1.1	1177.0 ± 2.2	194.6 ± 1.2
PN (ppbv)	0.53 ± 0.16	2.5 ± 0.2	0.43 ± 0.13

* Vertical offsets were modified in Fig. 5a to minimize overlap and improve clarity.

AMTD
7, 5953–6019, 2014

Table 2. Parameters of Eq. (1) fitted to the chromatograms in Fig. 5b, acquired during the FOSSILS 2013 field campaign. The fit uncertainties are those estimated using built-in algorithms in the software package Igor Pro 6. The uncertainty stated for PN is the 1σ measurement precision.

Parameter (unit)	PAN	PPN	MPAN
t_0 (s)	115.080 ± 0.002	243.820 ± 0.002	408.820 ± 0.002
σ (s)	1.667 ± 0.002	2.946 ± 0.002	4.912 ± 0.002
V_0 (mV)*	118.9 ± 2.1	64.3 ± 1.0	22.06 ± 0.34
m (mV s^{-1})	0.125 ± 0.018	0.0446 ± 0.004	0.0167 ± 0.0008
A (mV s)	1430.5 ± 1.8	1432.1 ± 1.0	1694.9 ± 0.7
PN (ppbv)	2.43 ± 0.20	2.66 ± 0.21	2.38 ± 0.19
Impurities (pptv)	n/a	PAN: 41	PAN: 610; PPN: 29

* Vertical offsets were modified in Fig. 5b to minimize overlap and improve clarity.

AMTD
7, 5953–6019, 2014

Table 3. Response factors of the PAN-GC (in mV s ppbv^{-1}) in the CAP and N_2 STD sensitivity setting and of the TD-CIMS (in Hz pptv^{-1}) for PAN and PPN during the Abbotsford campaign. TD-CIMS counts were normalized to 1×10^6 reagent ion counts at $m\ z$ 127. RF_u = uncorrected response factor, RF_{cor} = response factor corrected for PAN impurities contributing to PN signal, RSD = relative standard deviation, $r = r$ value of the calibration plot, n = number of calibration runs, n/d = not determined.

Parameter or reference	PAN (CAP)	PPN (CAP)	PAN (N_2 STD)	PPN (N_2 STD)	PAN (CIMS)	PPN (CIMS)
RF_u	761 ± 47	n/d	280 ± 5	272 ± 8	n/d	n/d
RF_{cor}	774 ± 42	n/d	281 ± 5	277 ± 10	23.3 ± 0.5	22.5 ± 1.0
RSD	5.4 %	n/d	1.7 %	3.6 %	2.1 %	4.4 %
r	0.985	n/d	0.997	0.994	0.993	0.989
n	6	–	11	6	16	7
relative response	1	n/d	1	0.99 ± 0.04	1	0.97 ± 0.05

AMTD
7, 5953–6019, 2014

Table 4. Response factors (in mV s ppbv^{-1}) of the PAN-GC for PAN, PPN, and MPAN during FOSSILS 2013 with the initial flow settings. Abbreviations are the same as in Table 3.

Parameter	PAN	PPN	MPAN
RF_u	633 ± 13	538 ± 24	178 ± 10
RF_{cor}	636 ± 11	576 ± 8	529 ± 27
RSD	1.7 %	1.4 %	5.1 %
r	0.993	0.997	0.962
n	29	16	17

AMTD
7, 5953–6019, 2014

Table 5. Response factors (in mV s ppbv^{-1}) of the PAN-GC for PAN, PPN, and MPAN during FOSSILS 2013, after the switch to a lower flow rate. Abbreviations are as in Table 3.

Parameter	PAN	PPN	MPAN
RF_u	943 ± 12	877 ± 13	624 ± 25
RF_{cor}	950 ± 11	886 ± 12	764 ± 15
RSD	1.2 %	1.4 %	2.0 %
r	0.999	0.999	0.997
n	5	5	9

AMTD
7, 5953–6019, 2014

Table 6. Response factors (in mV s ppbv^{-1}) of the PAN-GC for PAN, PPN, MPAN and APAN determined after the FOSSILS campaign. Abbreviations are as in Table 3.

Parameter	PAN	PPN	MPAN	APAN
RF_u	662 ± 24	637 ± 8	n/d	451 ± 13
RF_{cor}	662 ± 24	637 ± 8	n/d	(513 ± 13)*
RSD	3.6 %	1.3 %	n/d	2.5 %
r	0.990	0.9997	n/d	0.981
n	8	6	–	5

* Corrected for PAN only.

AMTD
7, 5953–6019, 2014

Table 7. Relative response factors of the PAN-GC for PAN, PPN, MPAN and APAN and recent literature values. nd = not determined; nr = determined, but not disclosed.

Reference	PAN	PPN	MPAN	APAN
Grosjean et al. (1993b)	1	0.73–0.81	nr	nd
Williams et al. (1993)	1	0.72 ± 0.15	0.21 ± 0.04	nd
Williams et al. (2000)	1	0.83	0.72	nd
Roberts et al. (2001)	1	0.83	nr	(0.92)*
Flocke et al. (2005)	1	0.90 ± 0.02	0.64 ± 0.03	(0.80 ± 0.15)*
Zheng et al. (2011)	1	nd	nd	0.959 ± 0.006
This work (Abbotsford)	1	0.99 ± 0.04	nd	nd
This work (FOSSILS 1)	1	0.91 ± 0.02	0.83 ± 0.04	nd
This work (FOSSILS 2)	1	0.93 ± 0.02	0.80 ± 0.02	nd
This work (Laboratory)	1	0.96 ± 0.04	nd	> (0.77 ± 0.03)**

* Estimated; ** corrected for PAN only.

AMTD
7, 5953–6019, 2014

Table 8. Response factors of the TD-CIMS for PAN, PPN, MPAN and APAN determined in the laboratory. Uncorrected response factors (RF_u), response factors corrected for sample impurities, when quantifiable (RF_{cor}), zero offsets (b), r value of the calibration plot, and response factors relative to PAN. TD-CIMS counts were normalized to 1×10^6 reagent ion counts at $m\ z\ 127$.

Parameter (unit)	PAN	PPN	MPAN	APAN
RF_u ($Hz\,pptv^{-1}$)	27.7 ± 0.3	24.7 ± 0.3	0.64 ± 0.02	16.5 ± 0.4
RF_{cor} ($Hz\,pptv^{-1}$)	27.7 ± 0.3	25.1 ± 0.3	0.8 ± 0.1	$> 20.9 \pm 0.4$
RSD	1.1 %	1.2 %	12.5 %	1.9 %
b (10^3Hz)	1.8 ± 0.6	3.1 ± 0.3	0.27 ± 0.05	2.0 ± 0.5
r	0.99	0.98	0.97	0.99

AMTD
7, 5953–6019, 2014

Table 9. Relative response factors of the TD-CIMS for PAN, PPN, MPAN and APAN in this work and recent literature values. n/d = not determined or reported.

Reference	PAN	PPN	MPAN	APAN
Slusher et al. (2004)	1	1	0.16	n/d
Turnipseed et al. (2006)	1	1	0.125	n/d
Wolfe et al. (2007)	1	1	0.125	1
LaFranchi et al. (2009); Wolfe et al. (2009)	1	1	0.232	n/d
Zheng et al. (2011)	1	1.13 ± 0.16	0.014 ± 0.002	0.66 ± 0.10 (cal source) 0.37 (FTIR in chamber) 0.35 ± 0.04 (recommended)
Mielke and Osthoff (2012)	1	1	0.022	n/d
This work	1	0.91 ± 0.01	0.029 ± 0.004	> (0.75 ± 0.02)*

* Corrected for PAN only.

AMTD
7, 5953–6019, 2014

Table 10. Parameters of Eq. (1) fitted to chromatograms in Figs. 8 and 9.

Parameter (unit)	PAN (Fig. 8)	PAN (Fig. 9)	PPN (Fig. 8)
t_0 (s)	119.894 ± 0.001	140.00 ± 0.01	239.76 ± 0.001
σ (s)	1.824 ± 0.001	2.03 ± 0.01	2.810 ± 0.001
V_0 (mV)	1.66 ± 0.02	13.97 ± 0.07	3.15 ± 0.03
m (V s^{-1})	$-(1.5 \pm 0.2) \times 10^{-5}$	$-(3.99 \pm 0.05) \times 10^{-5}$	$-(5.48 \pm 0.04) \times 10^{-5}$
A (mV s)	468.02 ± 0.25	6.26 ± 0.05	45.24 ± 0.03
RF_{cor} (mV s ppbv^{-1})	281 ± 5	774 ± 42	277 ± 10
Mixing ratio (pptv)	1670 ± 30	8.1 ± 0.4	163 ± 6

AMTD
7, 5953–6019, 2014

Table 11. Parameters of Eq. (1) fitted to chromatogram in Fig. 13.

Parameter (unit)	PAN	PPN	MPAN
t_0(s)	113.575 ± 0.005	245.12 ± 0.02	406.21 ± 0.03
σ (s)	2.083 ± 0.006	3.87 ± 0.02	5.060 ± 0.035
V_0(mV)	131.66 ± 0.15	128.54 ± 0.03	127.28 ± 0.02
m (V s^{-1})	$-(3.19 \pm 0.34) \times 10^{-5}$	$(1.81 \pm 0.29) \times 10^{-6}$	$(7.73 \pm 0.13) \times 10^{-6}$
A (mV s)	115.4 ± 0.3	12.04 ± 0.06	7.80 ± 0.06
RF_{cor} (mV s ppbv^{-1})	950 ± 11	886 ± 12	764 ± 15
Mixing ratio (pptv)	121.5 ± 1.5	13.6 ± 0.2	10.2 ± 0.3

AMTD
7, 5953–6019, 2014

Table 12. Parameters of Eq. (1) fitted to a selection of PAN peaks in the chromatograms acquired during the night of 2 to 3 September 2013. Data for the peaks shown in Fig. 14 are highlighted in bold font. The errors are those reported by the IGOR software, except for the PAN mixing ratio which was calculated by adding errors of the calibration uncertainty and the error in the peak area (reported by IGOR) in quadrature. Note that the fit of the peak in the 08:10 chromatogram did not converge.

Time of injection (UTC)	t_0 (s)	σ (s)	V_0 (mV)	m (10^{-6} V s^{-1})	A (mV s)	PAN (pptv)
06:00	110.13 ± 0.02	2.07 ± 0.02	0.12106 ± 0.00005	−22.5 ± 0.4	12.1 ± 0.1	12.7 ± 0.2
06:10	**110.41 ± 0.02**	**2.07 ± 0.02**	**0.12098 ± 0.00005**	**−25.4 ± 0.4**	**11.4 ± 0.1**	**12.0 ± 0.2**
06:20	110.53 ± 0.02	2.08 ± 0.02	0.12014 ± 0.00004	−23.5 ± 0.4	11.8 ± 0.1	12.4 ± 0.2
06:30	110.60 ± 0.02	2.09 ± 0.03	0.11873 ± 0.00005	−16.9 ± 0.4	8.9 ± 0.1	9.3 ± 0.2
06:40	109.98 ± 0.02	2.08 ± 0.02	0.11866 ± 0.00005	−20.5 ± 0.4	8.9 ± 0.1	9.3 ± 0.1
06:50	110.07 ± 0.03	2.09 ± 0.03	0.11849 ± 0.00004	−23.4 ± 0.4	6.2 ± 0.1	6.5 ± 0.1
07:00	109.72 ± 0.03	2.08 ± 0.03	0.11771 ± 0.00005	−20.6 ± 0.5	7.3 ± 0.1	7.7 ± 0.1
07:10	108.41 ± 0.09	2.09 ± 0.09	0.11769 ± 0.00004	−23.6 ± 0.4	2.4 ± 0.1	2.6 ± 0.1
07:20	109.89 ± 0.02	2.08 ± 0.02	0.11770 ± 0.00005	−25.8 ± 0.4	9.1 ± 0.1	9.6 ± 0.2
07:30	108.96 ± 0.07	2.09 ± 0.07	0.11693 ± 0.00004	−22.4 ± 0.3	2.6 ± 0.1	2.8 ± 0.1
07:40	**108.35 ± 0.04**	**2.08 ± 0.05**	**0.11656 ± 0.00004**	**−22.0 ± 0.4**	**3.6 ± 0.1**	**3.8 ± 0.1**
07:50	108.55 ± 0.08	2.08*	0.11648 ± 0.00005	−24.2 ± 0.4	2.7 ± 0.1	2.8 ± 0.1
08:00	108.80 ± 0.08	2.08*	0.11600 ± 0.00004	−23.2 ± 0.4	2.7 ± 0.1	2.8 ± 0.1
08:10	–	–	–	–	–	–
08:20	109.22 ± 0.07	2.09 ± 0.07	0.11578 ± 0.00004	−23.0 ± 0.4	2.8 ± 0.1	2.9 ± 0.1
08:30	108.21 ± 0.06	2.09*	0.11578 ± 0.00004	−24.4 ± 0.4	2.0 ± 0.1	2.1 ± 0.1
08:40	107.17 ± 0.17	2.09*	0.11559 ± 0.00004	−24.2 ± 0.4	1.5 ± 0.1	1.6 ± 0.2
08:50	108.20 ± 0.10	2.08 ± 0.10	0.11538 ± 0.00004	−24.3 ± 0.4	1.6 ± 0.1	1.7 ± 0.1
09:00	106.70 ± 0.10	2.08*	0.11536 ± 0.00003	−26.3 ± 0.3	1.7 ± 0.1	1.8 ± 0.1
09:10	107.68 ± 0.07	2.08*	0.11535 ± 0.00005	−27.3 ± 0.4	1.6 ± 0.1	1.7 ± 0.1
09:20	**106.80 ± 0.15**	**2.09 ± 0.16**	**0.11447 ± 0.00004**	**−21.3 ± 0.3**	**1.1 ± 0.1**	**1.1 ± 0.1**
09:30	106.59 ± 0.13	2.08 ± 0.14	0.11453 ± 0.00004	−22.0 ± 0.4	1.1 ± 0.1	1.1 ± 0.1
09:40	106.90 ± 0.11	2.07 ± 0.12	0.11451 ± 0.00004	−23.3 ± 0.4	1.5 ± 0.1	1.6 ± 0.1
09:50	106.46 ± 0.17	2.07*	0.11429 ± 0.00004	−21.8 ± 0.4	1.5 ± 0.1	1.6 ± 0.1
10:00	104.45 ± 0.17	2.07*	0.11429 ± 0.00004	−22.6 ± 0.4	1.6 ± 0.1	1.6 ± 0.1
10:10	106.90 ± 0.11	2.06 ± 0.12	0.11439 ± 0.00005	−24.5 ± 0.5	1.8 ± 0.1	1.9 ± 0.1
10:20	107.30 ± 0.09	2.07 ± 0.09	0.11429 ± 0.00004	−24.3 ± 0.4	1.9 ± 0.1	2.0 ± 0.1

* σ constrained to value from preceding run.

AMTD
7, 5953–6019, 2014

Table 13. Selected LODs of ambient air measurements of PAN, PPN and MPAN by GC-ECD reported in the literature. nr = not reported, nd = not determined.

Reference	Measurement location	PAN	PPN	MPAN
Grosjean et al. (1993b)	California mountains	10	20	200
Williams et al. (1993)	Atlanta, Georgia	30	nr	100–150
Nouaime et al. (1998)	Tennessee		2–5	
Williams et al. (2000)	Tennessee		< 1–5	
Thornberry et al. (2001)	Michigan	2	3	5
Roberts et al. (2002)	Tennessee	5	5	5
Roberts et al. (2006)	New England	4	4	4
This work	Abbotsford, British Columbia	4	16	nd
This work	Fort McKay, Alberta	1	2	3

AMTD
7, 5953–6019, 2014

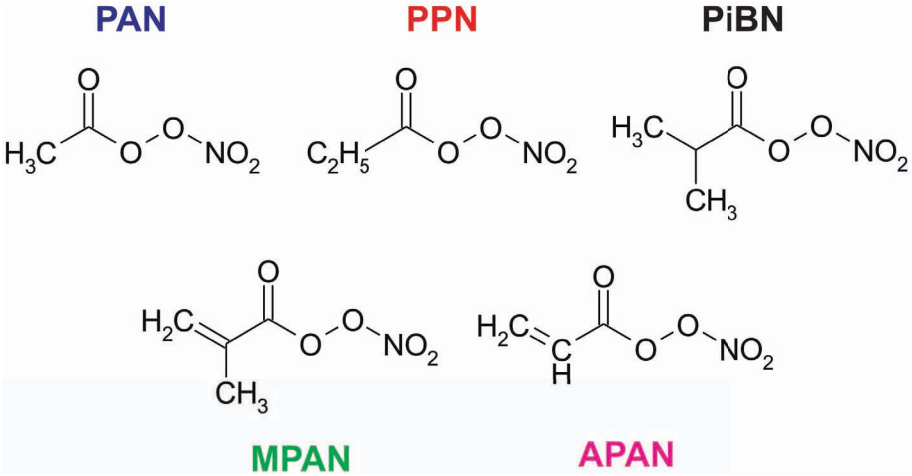

Figure 1. Molecular structures of common peroxycarboxylic nitric anhydrides. PAN = peroxyacetic, PPN = peroxypropionic, PiBN = peroxyisobutanoic, MPAN = peroxymethacryloyl, and APAN = peroxyacryloyl nitric anhydride.

AMTD
7, 5953–6019, 2014

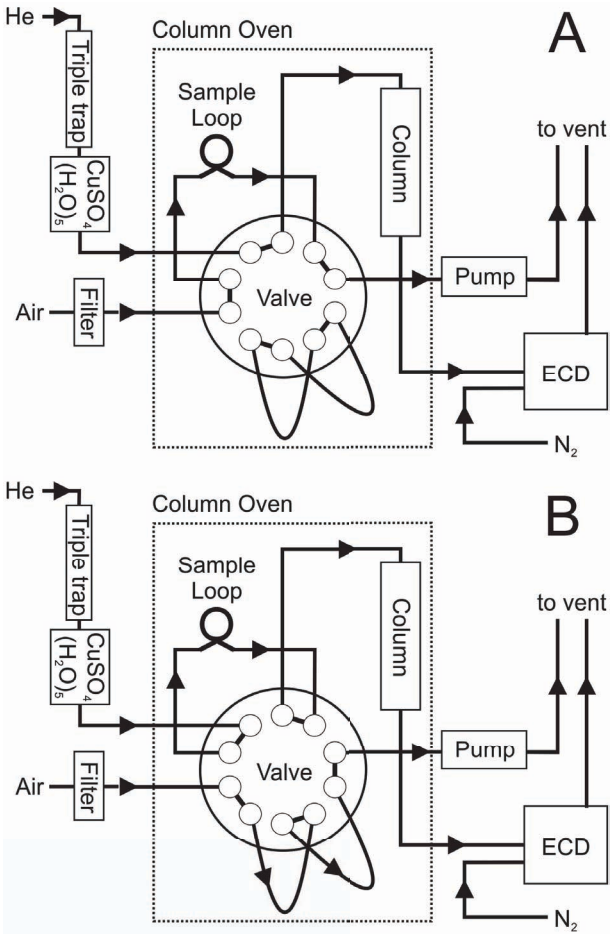

Figure 2. Schematic of the PAN-GC (items not to scale). **(A)** Fill sample loop (default). **(B)** Sample loop to column. See text for a detailed description.

AMTD
7, 5953–6019, 2014

A gas chromatograph for quantification of PANs

T. W. Tokarek et al.

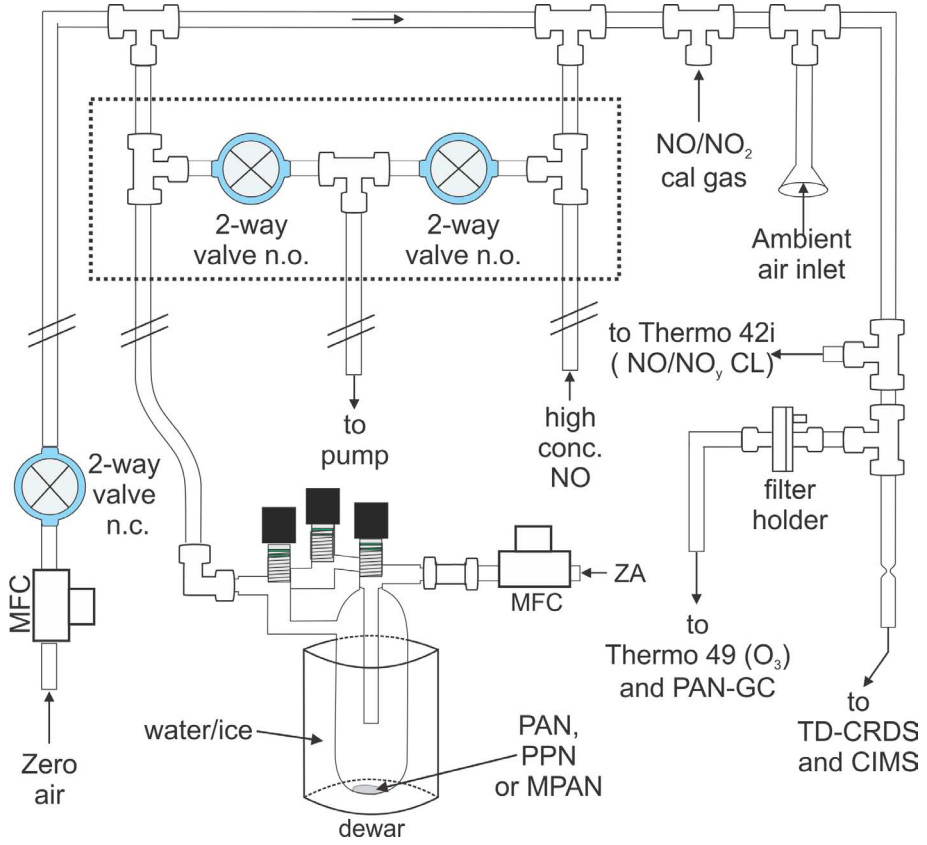

Figure 3. Schematic of the inlet and calibration setup used during the Abbotsford 2012 and FOSSILS 2013 field campaigns. MFC = mass flow controller, CL = Chemiluminescence, n.o. = normally open, n.c. = normally closed.

AMTD
7, 5953–6019, 2014

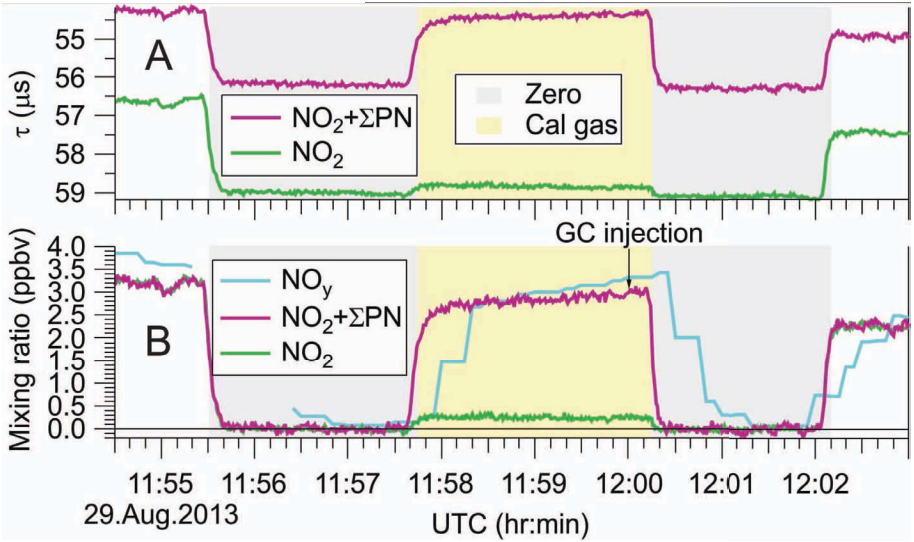

Figure 4. Typical calibration sequence. **(A)** Time series of ring-down time constants. **(B)** Time series of TD-CRDS and NO_y mixing ratios. The instruments were "zeroed" at ~ 11:55:30 UTC. PPN was added at ~ 11:57:40 UTC. The GC injected at noon (the chromatogram is shown in Fig. 5b). Immediately after the injection, the instruments were zeroed again. The NO_y signal was not adjusted for its slow response to concentration changes in this Figure.

AMTD
7, 5953–6019, 2014

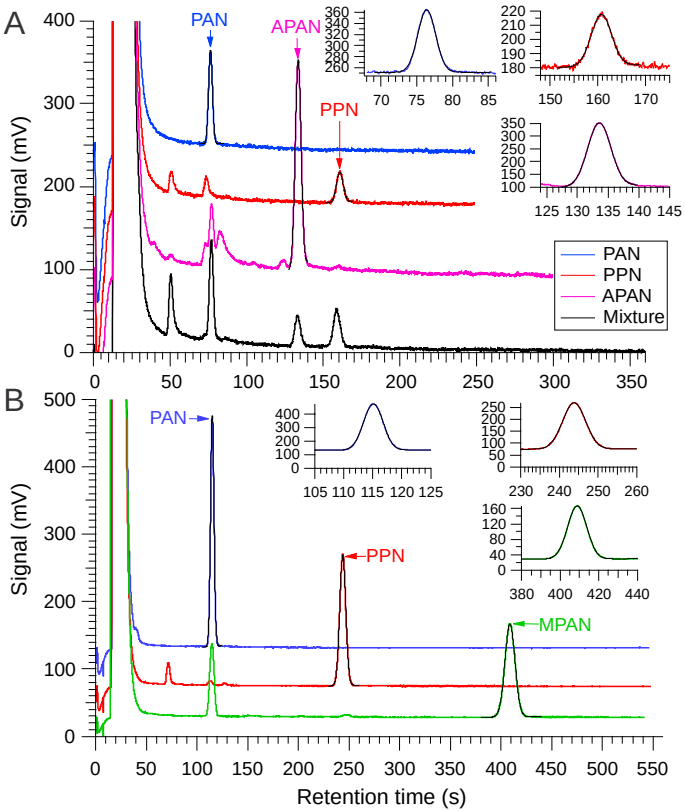

Figure 5. Sample chromatograms (vertically offset for clarity) of PAN, PPN, APAN, and MPAN standards on a 10.4 m RTX-1701 column at a He flow rate of 20 mL min^{-1}. **(A)** Chromatograms of standards acquired in the laboratory at a column oven temperature of 27 °C. **(B)** Chromatograms of standards acquired during the FOSSILS 2013 field study at an oven temperature of 22 °C. The fits of expression Eq. (1) to the observed peaks are superimposed as black traces.

AMTD
7, 5953–6019, 2014

A gas chromatograph for quantification of PANs

T. W. Tokarek et al.

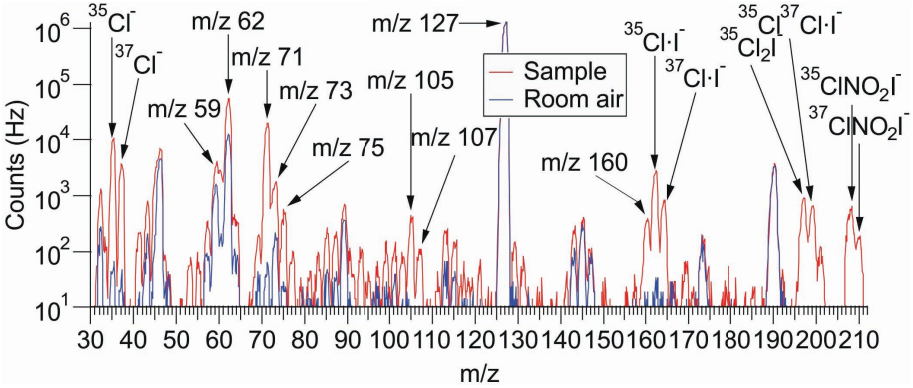

Figure 6. Iodide ion chemical ionization mass spectra of an APAN standard and room air.

AMTD
7, 5953–6019, 2014

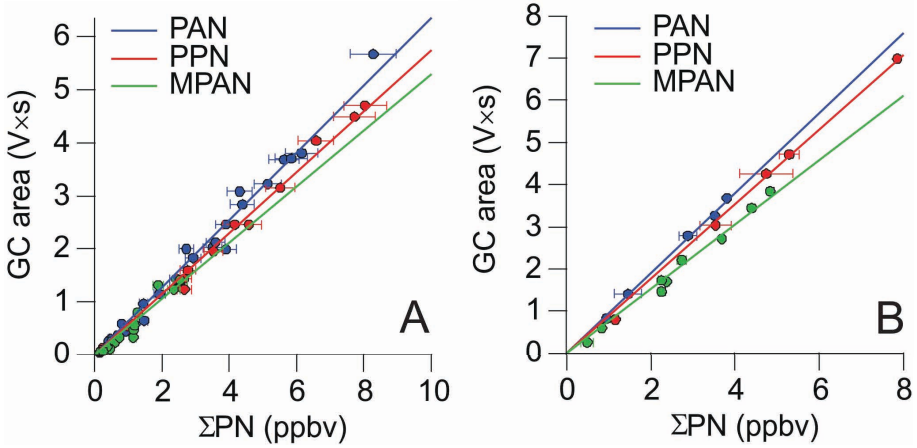

Figure 7. Calibration plots of PAN-GC vs. TD-CRDS during FOSSILS 2013. (**A**, left) Data collected with the initial sampling conditions. (**B**, right) Data collected after change of sample flow rate due to pump failure. The error bars indicate the precision of the measurements ($\pm 1\sigma$) that include instability of the source output. See Tables 3 and 4 for the linear regression statistics.

AMTD
7, 5953–6019, 2014

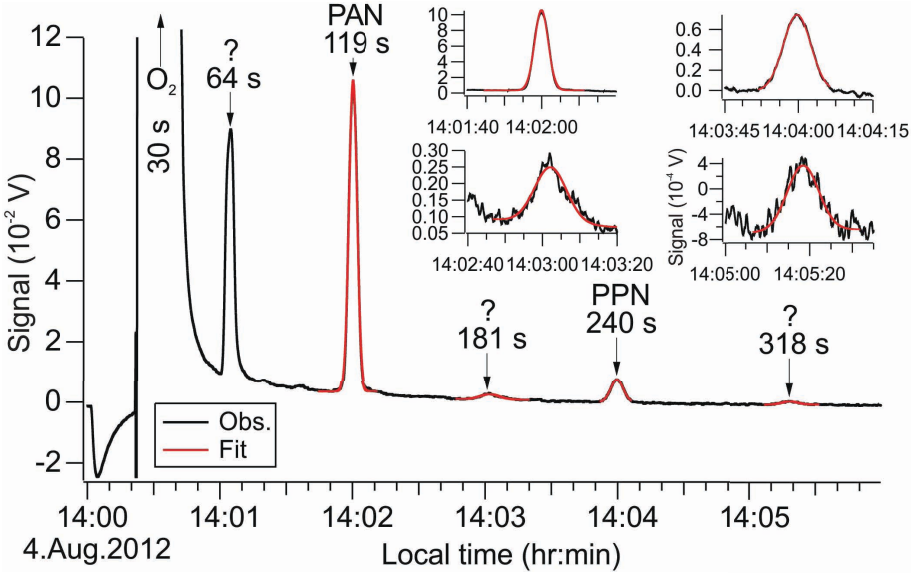

Figure 8. Chromatogram of ambient air acquired in Abbotsford on 4 August 2012, 14:00 LT using a 20.41 m RTX-1701 column at a He flow rate of 20.0 mL min^{-1} with the PAN-GC operated in "N_2STD" setting mode. The chromatogram was smoothed before presentation (using Igor Pro's Savitzky-Golay algorithm, 4th order, 1001 data points, ~2 s). The red lines are fits of Eq. (1) to the observed peaks, which are also shown as inserts. The peak areas correspond to a PAN mixing ratio of 1.67 ppbv and a PPN mixing ratio of 163 pptv, which were the largest peak areas observed in Abbotsford.

AMTD
7, 5953–6019, 2014

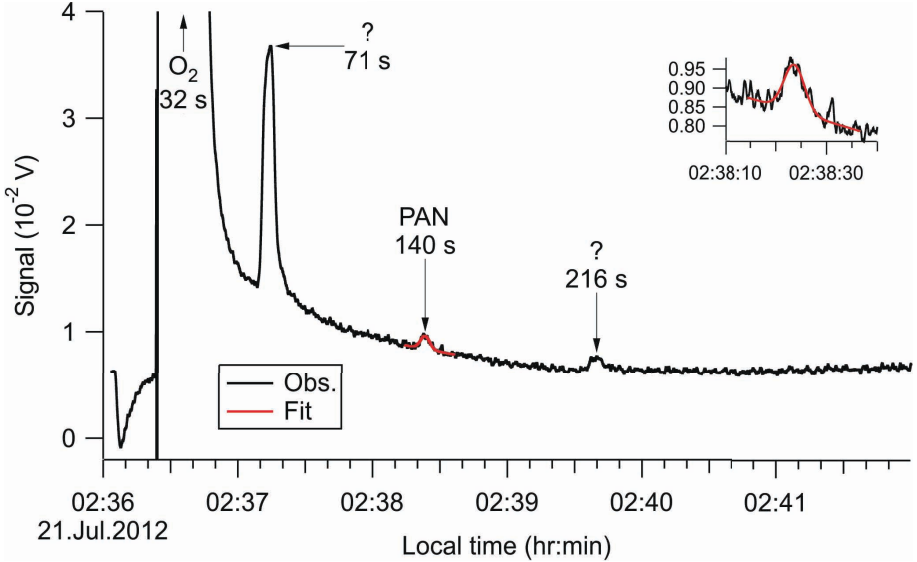

Figure 9. Chromatogram of ambient air acquired in Abbotsford on 21 July 2012, 02:36 local time with the PAN-GC operated in "CAP" setting mode. Other conditions were the same as in Fig. 8. The peak area corresponds to a PAN mixing ratio of (8.1 ± 0.4) pptv. Retention times are greater than in Fig. 8 due to a lower column temperature.

AMTD
7, 5953–6019, 2014

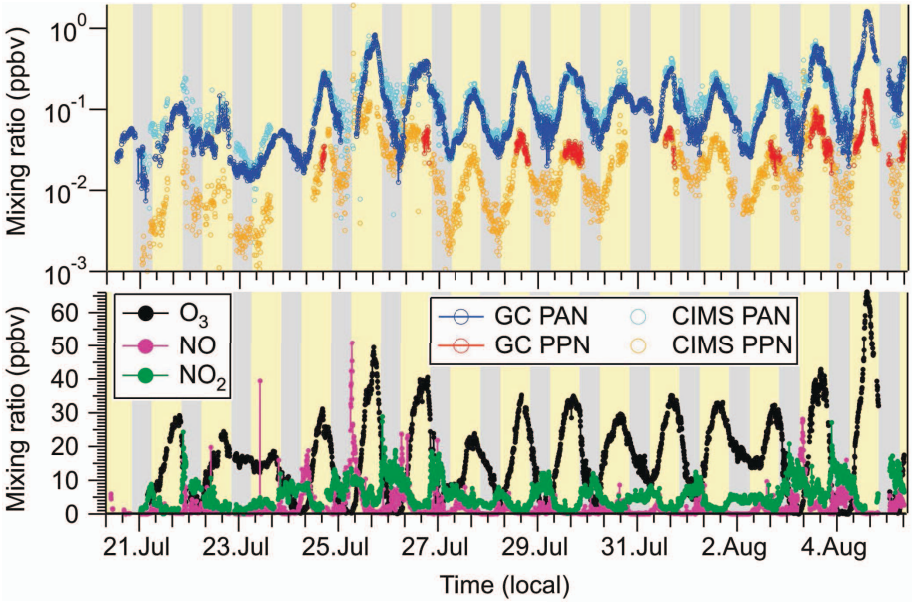

Figure 10. Time series of PAN and PPN (top) and NO, NO_2, and O_3 mixing ratios (bottom) observed in Abbotsford.

AMTD
7, 5953–6019, 2014

A gas chromatograph for quantification of PANs

T. W. Tokarek et al.

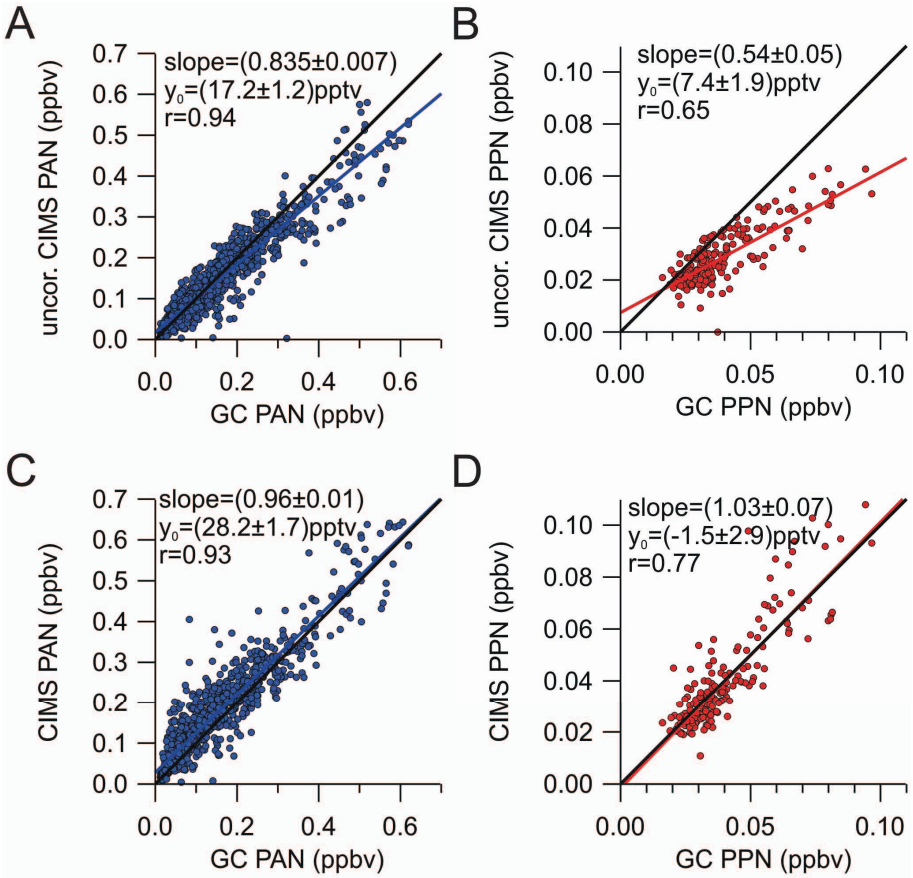

Figure 11. Scatter plots of TD-CIMS and PAN-GC data. **(A)** PAN, uncorrected. **(B)** PPN, uncorrected. **(C)** PAN, corrected. **(D)** PPN, corrected.

AMTD
7, 5953–6019, 2014

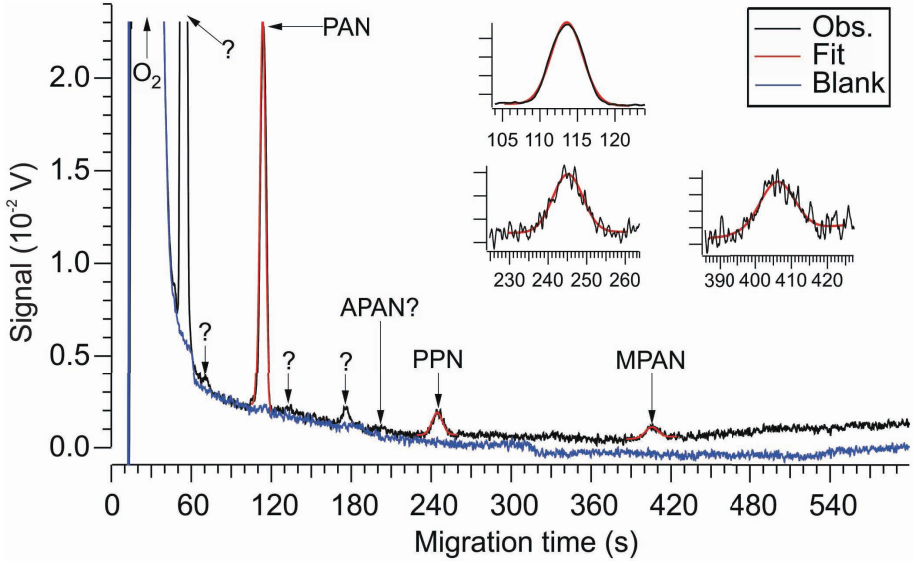

Figure 12. Chromatogram of ambient air acquired during FOSSILS on 4 September 2013, 15:30 LT and of zero air blank acquired at 17:30 local time using a 10.4 m RTX-1701 column at a He flow rate of 20.0 mL min^{-1}. The chromatograms were smoothed before presentation (Algorithm: Savitzky-Golay, 4th order, 41 data points, ~ 2 s). The red lines are fits of (1) to the observed peaks, which are superimposed as inserts for clarity. The peak areas correspond to the following mixing ratios: PAN (121.5±1.5) pptv, PPN: (13.6±0.2) pptv, and MPAN: (10.2±0.3) pptv.

AMTD
7, 5953–6019, 2014

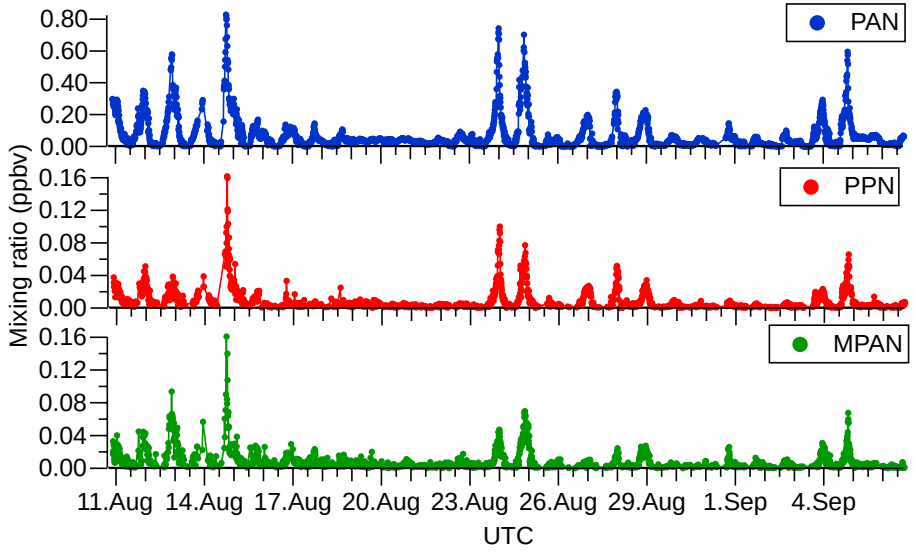

Figure 13. Time series of PAN, PPN, an MPAN during FOSSILS.

AMTD
7, 5953–6019, 2014

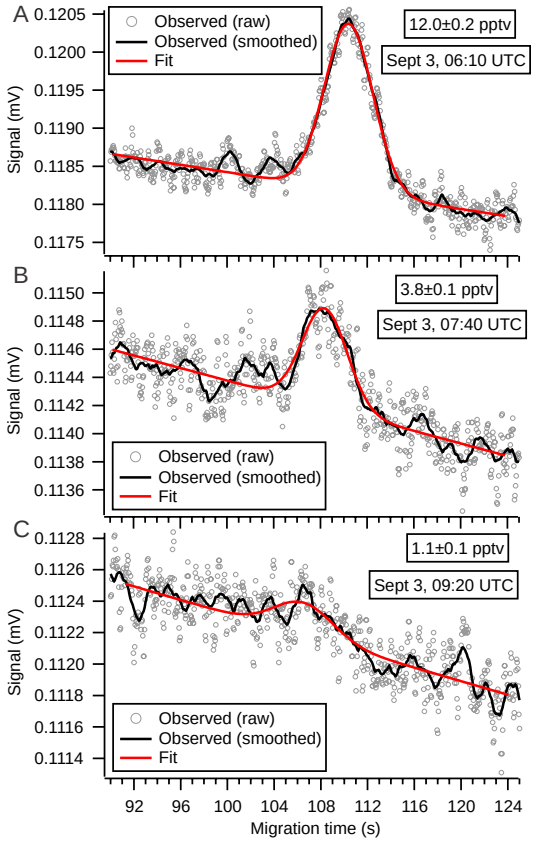

Figure 14. Chromatograms of PAN acquired on 3 September 2013. The fit to the peaks in **(A)** and **(B)** were unconstrained; the fit to the peak in **(C)** was constrained in its width parameter (σ) to the earlier determined value (at higher PAN concentration) of 2.1 s. The mixing ratios are those of the peak area of the fitted line (shown in red).

AMTD
7, 5953–6019, 2014

A gas chromatograph for quantification of PANs

T. W. Tokarek et al.

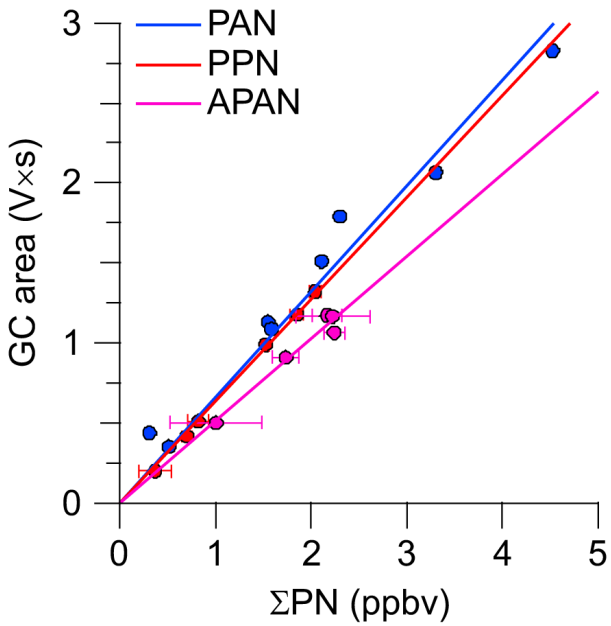

Figure A1. Stability of the PAN-GC retention times (top) and relative peak widths (bottom). The temperature inside the trailer was considerably more stable after 21 August, after an auxiliary air conditioning unit was installed.

AMTD
7, 5953–6019, 2014

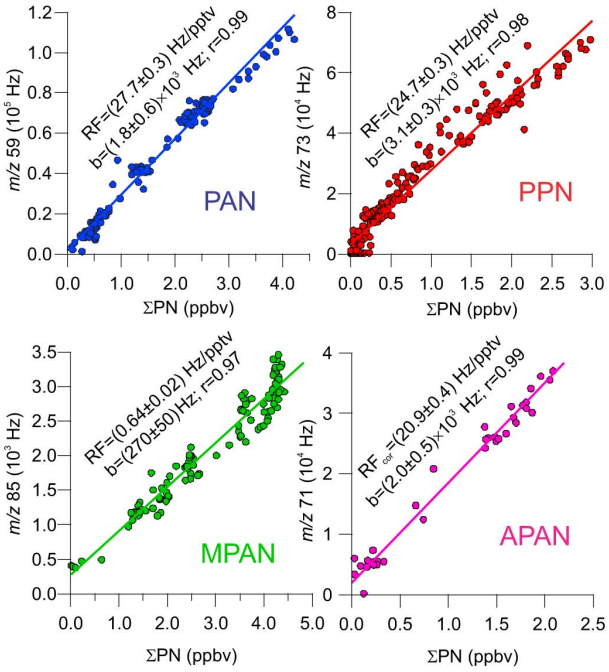

Figure A2. PAN-GC calibration plots for PAN, PPN, and APAN (post-FOSSILS).

AMTD
7, 5953–6019, 2014

A gas chromatograph for quantification of PANs

T. W. Tokarek et al.

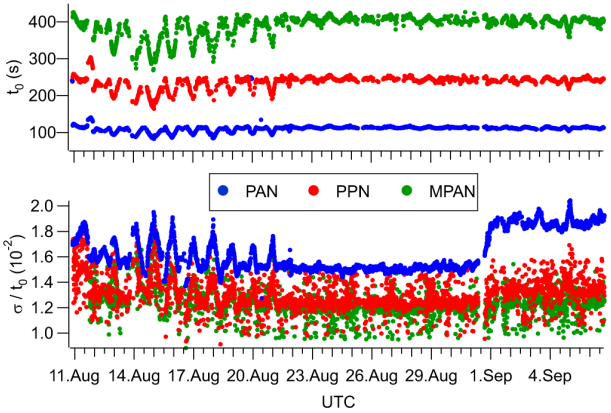

Figure A3. TD-CIMS calibration plots for PAN, PPN, MPAN, and APAN.

AMTD
7, 5953–6019, 2014

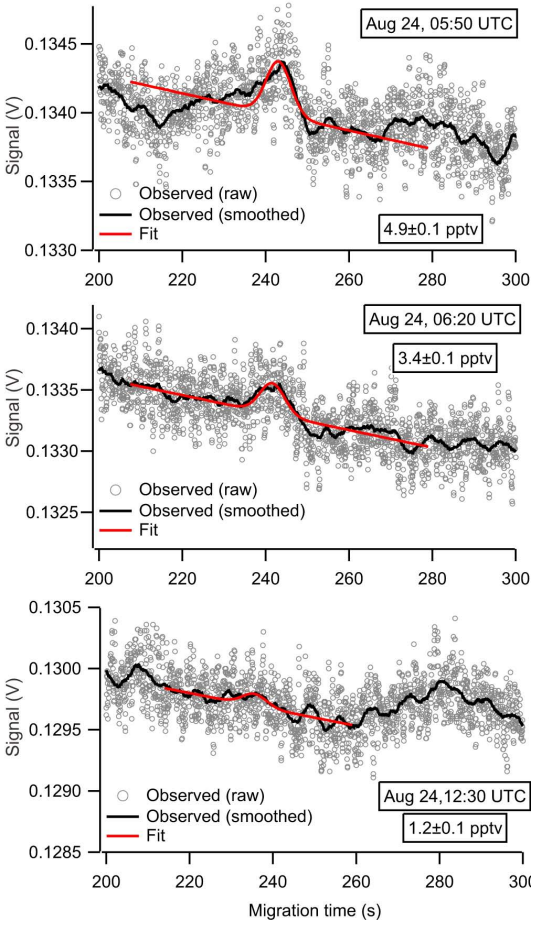

Figure A4. Sample chromatograms of PPN acquired on 24 August 2013 (analogous to those in Fig. 14) used to estimate limit of detection for PPN.

AMTD
7, 5953–6019, 2014

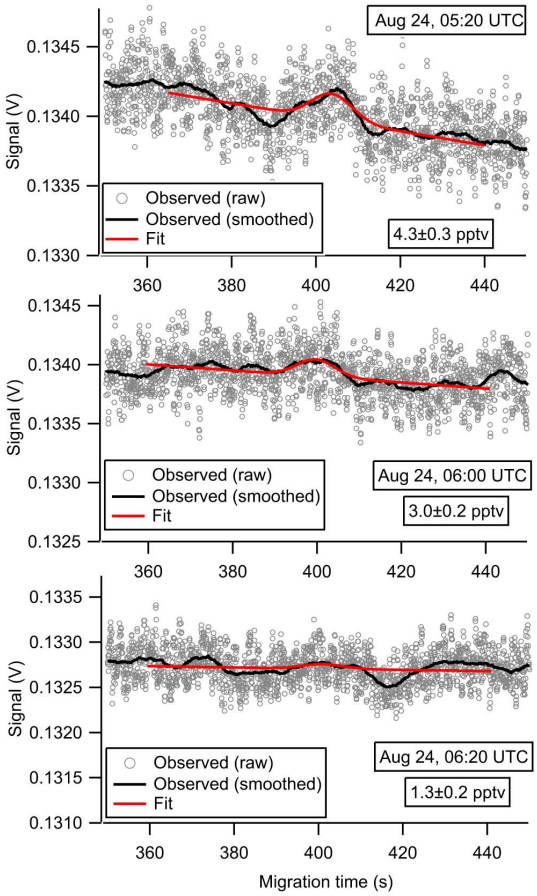

Figure A5. Sample chromatograms of MPAN acquired on 24 August 2013 (analogous to those in Figs. A4 and 14) used to estimate limit of detection for MPAN.

AMTD
7, 5953–6019, 2014